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1 **Shifts in the elemental composition of plants during a very severe drought**2 **Ifigenia Urbina^{1,2}, Jordi Sardans^{1,2*}, Carl Beierkuhnlein³, Anke Jentsch⁴, Sabrina**
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5 Spain.6 ²CREAF, 08913 Cerdanyola del Vallès, Catalonia, Spain.7 ³Biogeography, BayCEER, University of Bayreuth, 95440 Bayreuth, Germany.8 ⁴Disturbance Ecology, BayCEER, University of Bayreuth, 95440 Bayreuth, Germany.9 *Corresponding author. Tel. 34 93 581 2934. Fax: 34 93 581 4151. E-mail address:
10 j.sardans@creaf.uab.es.11 **Highlights**

- 12 • -Each species had a particular elemental composition in all circumstances
- 13
- 14 • -Species had more plastic stoichiometry when growing in more diverse communities
- 15
- 16 • -The stoichiometric shifts due to drought were more community- than species-
17 dependent
- 18
- 19 • -Drought in all cases decreased K, N, P, Mg and S, and increased C and Fe
20 concentrations

21 **ABSTRACT**

22 Diverse plant functions (e.g. growth, storage, defense and anti-stress mechanisms) use
23 elements disproportionately. We hypothesized that plants growing under different

24 abiotic and biotic conditions would shift their elemental compositions in response to a
25 very severe drought. We tested this hypothesis by investigating the changes in foliar
26 stoichiometry and species composition from a very severe drought. We also tested the
27 effects of previous droughts (acclimation) on this response. Different species growing
28 in the same community responded more similarly to a very severe drought than did
29 individual species growing in different communities. The stoichiometric shifts were
30 thus more community-dependent than species-dependent. The results also suggested
31 that plants grown in monoculture were less stoichiometrically plastic during the
32 drought than plants grown in a more diverse community. Previous exposure to long-
33 term drought treatments in the same communities did not significantly affect the
34 stoichiometric shifts during the new drought. Differential use of resources may have
35 been responsible for these responses. Monocultured plants, which used the same
36 resources in similar proportions, had more difficulty avoiding direct competition when
37 the resources became scarcer. Moreover, each species tested had a particular
38 elemental composition in all communities and climatic treatments. The differences in
39 foliar elemental compositions were largest between plant functional groups (shrubs
40 and grasses) and smallest among species within the same functional group. Global
41 principal components analyses (PCAs) identified a general tendency for all species,
42 independently of the community in which they grew, toward lower concentrations of
43 K, N, P, Mg and S, [and to higher concentrations of C and Fe as the drought advanced](#).
44 This study has demonstrated the utility of analyses of differences and shifts in plant
45 elemental composition for understanding the processes underlying the responses of
46 plants to changes in biotic and abiotic environmental conditions.

47 Keywords: Biogeochemical niche; Ecological stoichiometry; elemental composition;
48 extreme drought; nitrogen; phosphorus.

49 **1. Introduction**

50 Climate change is modifying the patterns of global precipitation and the intensity and
51 frequency of droughts (IPCC, 2013), with important consequences to plant physiology,
52 community diversity and ecosystemic productivity in terrestrial biomes (Allen et al.,
53 2010; Trnak et al., 2010; McDowell et al., 2011; Weiβhuhn et al., 2011; IPCC., 2013;
54 Peñuelas et al., 2013). The capacity to shift elemental compositions in response to
55 diverse types of drought stress could therefore be expected to vary among plant
56 species.

57 The main goal of ecological stoichiometry is to understand the relationships of
58 environments and C:N:P ratios with the lifestyles of organisms and the structures and
59 functions of ecosystems (Sterner and Elser, 2002; Elser and Hamilton, 2007). These
60 relationships have been widely observed in planktonic communities, where low N:P
61 ratios in the medium are correlated with high growth rates in organisms and are
62 consequently important to the structure of trophic webs (Elser et al., 1996; Peñuelas et
63 al., 2013). Terrestrial plants, however, are larger than planktonic organisms, and N and
64 P can be allocated to functions other than growth, such as storage, defense or anti-
65 stress systems (Peñuelas and Sardans, 2009). Moreover, the concentrations or ratios of
66 other elements such as K, Mg, Ca and Fe in terrestrial plant communities are much
67 more involved in important physiological functions of organisms (Wilkinson et al.,
68 1990; Leung et al., 1994; Lombini et al., 2003; Murren et al., 2006; Wu et al., 2009;
69 Lourido et al., 2010; Hashimoto and Kudla, 2011; Sardans et al., 2012a, 2012b; Sardans

70 and Peñuelas, 2012; Stael et al., 2012; Yang et al., 2012; Perea-García et al., 2013) and
71 consequently should also be included in ecological stoichiometric studies. The
72 biogeochemical niche hypothesis claims that different sympatric plant species occupy
73 different biogeochemical niches, as manifested by the complete spectra of the
74 contents of macro- and micronutrients in their tissues (Peñuelas et al., 2008, 2010) and
75 not only by N and P concentrations or N:P ratios. The elemental composition of a
76 species should thus be determined, in part, by long-term genetic adaptation to a
77 particular abiotic environment and also by specific ecological strategies such as the
78 optimal successional stage. The success of a plant species, however, requires some
79 degree of flexibility to allow plants to respond to abiotic and biotic shifts.
80 Stoichiometric flexibility enables organisms to adjust their elemental ratios while
81 maintaining a constant function (Sistla and Schimel, 2012). Species with higher
82 stoichiometric flexibility tend to have higher concentrations of N and P and lower N:P
83 ratio (Yu et al., 2011).

84 K is strongly correlated with several physiological functions in plants such as
85 water-use efficiency and stomatal control and with above- and belowground biomass
86 and likely has an important role in photosynthesis (Leigh and Wyn Jones, 1984;
87 Andersen et al., 1992; Egilla et al., 2001; Cakmak, 2005, Fernandez et al., 2006, Sardans
88 et al., 2012c). Some studies have observed higher foliar K contents in drier periods,
89 especially in plants typical of dry environments (Rivas-Ubach et al., 2012; Sardans et
90 al., 2012c; 2013). Increases in C:nutrient ratios in leaves due to increases in C-rich
91 compounds linked to the avoidance of water stress are well-known stoichiometric
92 responses to decreases in water availability, at least in typically water-stressed
93 environments (Gonçalves-Alvim et al., 2004 and 2006; Sardans and Peñuelas, 2013).

94 The production of more sclerophyllous leaves and a slow growth rate are physiological
95 responses to high C:nutrient ratios, drought conditions and dry environments (Feller,
96 1995; Fernández and Reynolds, 2000; Gonçalves-Alvim et al., 2004 and 2006; Sardans
97 et al., 2011). The various possible responses to water stress depend on the dominant
98 mechanism of water uptake (isohydric vs anisohydric), the functional type, the species,
99 the soil nutrient resource and the nature of episodes of water stress (duration or
100 frequency) (DeLucia and Schlesinger, 1991; Tradieu and Simonneau, 1998; Schultz,
101 2003).

102 The frequency and magnitude of extreme climatic events such as drought will
103 increase as the global climate changes (IPCC, 2012). Several specific responses to
104 extreme droughts at the community level, however, remain to be studied. Water
105 shortage should lead to a decline in water potential and to water stress. Plants should
106 be able to alter their functions to increase the efficiencies of their mechanisms of
107 water uptake and use efficiency. These adaptations should depend on the composition
108 of the community, with heathlands being very resistant (Gordon et al., 1999; Peñuelas
109 et al., 2004) and grassland communities being more sensitive to extreme droughts
110 (Grime et al., 2000; Kreyling et al., 2008a). The capacity to adapt to and resist drought
111 should be accompanied by stoichiometric changes that further allow changes in the
112 allocation of elements to different functions and structures. A species growing in
113 communities with different species compositions should have distinct capacities to
114 access different nutrients, thereby changing its capacity to respond to extreme
115 droughts. We hypothesized that changes in the availability of water will be reflected in
116 changes in foliar stoichiometry, which in turn will ultimately help us to understand the
117 processes involved in the response to water stress. This response can also vary due to

118 the biotic relationships of plants. Previous studies have shown that plants of the same
119 species exhibit different elemental compositions depending on the species
120 composition of the plant community in which they live (Urbina et al., 2014).

121 Several studies have observed that plants subjected to recurrent extreme
122 climatic events are better able to cope with subsequent extreme events (Bruce et al.,
123 2007; Lambers et al., 2008; Walter et al., 2011). Such a capacity for ecological
124 "memory", though, has not been consistently observed (Zavalloni et al., 2008).
125 Modified responses after pre-exposure to extreme climatic events are also dependent
126 on the type of impact, and the opposite effect, worse performance, can also occur,
127 mainly in perennial plants; extreme frosts (in timing or intensity) can hurt perennial
128 plants and substantially reduce their vitality. Lloret et al. (2004) observed that
129 reproduction and survival were lower in *Quercus ilex* trees after a second extreme
130 drought in 1995 than after an earlier drought in 1985. The response of plants to an
131 extreme drought can thus also be negatively influenced by previous exposure.
132 Stoichiometry may be a key to understanding the underlying physiological processes
133 such as resource depletion or hormonal regulation. We thus aimed to determine if
134 plants subjected to previous recurrent periods of drought had different stoichiometric
135 responses to a new extreme drought.

136 We analyzed four typical central European grass and shrub communities to
137 investigate: (i) the species-specific foliar stoichiometric changes in response to an
138 extreme drought, (ii) the role of community composition in plant responses and (iii)
139 the role of previous droughts (acclimation) in these responses.

140

141

142 **2. Materials and Methods**143 *2.1. Experimental area and design*

144 This research is part of the ongoing EVENT I experiment (Jentsch et al., 2007, 2011) at
145 the Ecological-Botanical Garden of the University of Bayreuth, Germany (49°55'18"N,
146 11° 34'55"E; 365 m a.s.l.) that studies the effects of extreme climatic events. The
147 experiment focuses on important perennial temperate plant communities. Mean
148 annual air temperature at the site is 8.2 °C. Mean annual precipitation is 724 mm and
149 is distributed bi-modally with a major peak in June/July and a secondary peak in
150 December/January (data from the German Weather Service).

151 Our experiment had three fully crossed factors: (1) recurrent drought over five
152 years (drought and control), (2) four community compositions (two species of one
153 functional group, four species of two functional groups, four species of three
154 functional groups and four of these species in monocultures; see Section 2.3) and (3)
155 duration of a very severe drought for all plots, irrespective of their previous exposure
156 to drought. Three levels of duration were tested by sampling at three times during the
157 three-month drought (at two, six and twelve weeks after the onset of the drought).

158 The experimental setup consisted of five replicates of each factorial
159 combination, with a total of 30 2×2 m plots for the two species of dwarf shrubs (Table
160 1) (two pre-exposed treatments and three community compositions) and a total of 40
161 2×2 m plots for the grass species (two pre-exposure treatments and two community
162 compositions). The factors were applied in a split-plot design with the community

163 compositions blocked and randomly assigned within each of the five replicates of the
164 pre-exposure treatments. All communities grew in soil 80 cm in depth, including 20 cm
165 of topsoil, sourced from a nearby sand quarry. The topsoil had a total carbon content
166 of 2% and a pH of 4.5 (measured in 1 M KCl), and the lower soil layer had a total
167 carbon content of 0.2% and a pH of 6.2. The texture of the soil was loamy sand (82%
168 sand, 13% silt and 5% clay). The total concentrations of the various elements in the soil
169 are provided in Table S1.

170 *2.2. Recurrent and final very severe droughts*

171 The climatic manipulations consisted of two phases of drought, recurrent droughts
172 over five years and a very severe drought of three months in the final year of the
173 experiment. The controls were exposed to ambient conditions. The severity of the
174 droughts was determined by statistical extremity in a historical reference period
175 (extreme value theory) independent of its effects on organisms (Jentsch et al. 2007). In
176 particular, the intensity of the treatments was based on the local 100-year extreme
177 droughts in 2005, 2006 and 2007 and on the local 1000-year extreme droughts in
178 2008, 2009 and 2010. Gumbel distributions (Gumbel, 1958) were fitted to the annual
179 extremes, and the 100-year and 1000-year recurrent droughts were calculated.
180 Drought was defined as the number of consecutive days with less than 1 mm of daily
181 precipitation. Accordingly, a recurrent drought of 32 days (2005–2007) and of 42 days
182 (2008, 2009 and 2010) was applied during the peak growing season in June. The very
183 severe drought was applied in all plots (control and drought) from 17 May to 28 August
184 2011.

185 The severe drought was effected by large rain-out shelters (Haygrove Tunnels
186 Ltd., Ledbury, United Kingdom; total area 50 m in length × 31.5 m in width × 3.75 m in
187 height) over the entire surface of the experiment that permitted nearly 90%
188 transmission of photosynthetically active radiation through a transparent polyethylene
189 sheet 0.18 mm in thickness (UV M 42; folitec Agrarfolien-Vertriebs GmbH, Westerburg,
190 Germany). Unwanted greenhouse effects were avoided by leaving gaps between the
191 ground and the shelters 1.05 m along the sides and 2.55 m on the front and back,
192 allowing for near-surface air exchange. Before drought manipulation began, the plants
193 were watered with a total of 46.6 mm of water in two applications (11 and 13 May
194 2011) to adjust all plants to the same initial conditions. The amount of 46.6 mm was
195 the difference between the long-term average precipitation for April and the amount
196 of natural rainfall for April 2011 (from the start of the growing season to the start of
197 the severe drought). Lateral surface runoff was prevented by plastic sheets around the
198 treated plots buried to a depth of 10 cm. This study focused on the effects of the very
199 severe drought.

200 This drought consisted of nearly 3.5 months of no precipitation, which was
201 outside the calculated probabilities based on recorded weather data. Defining
202 "extreme" climates based on historical data, however, has limited value, and the
203 sensitivities of plant species to the same event may differ substantially (Hegerl et al.,
204 2011). The length of the drought in this experiment was not intended to represent a
205 certain future likelihood of duration but was rather considered as an experimental tool
206 to identify plant responses during the ongoing episode of drought (Kreyling et al.,
207 2014).

208

209 2.3. Experimental communities – composition and diversity

210 Four levels of diversity with different compositions of grasses, shrubs and herbs were
211 tested. The species are widespread in central Europe and are of fundamental
212 importance for agriculture and nature conservation (Table 1) (Jentsch *et al.* 2007,
213 2011). All species are perennials and have life spans of several years. The compositions
214 of the communities were maintained by periodic weeding three times per year. The
215 four community compositions were: (1) H2, composed of two dwarf shrubs of the
216 same functional group, *Calluna vulgaris* and *Vaccinium myrtillus*; (2) G4, composed of
217 two grasses of the same functional group, *Arrhenatherum elatius* and *Holcus lanatus*,
218 and two forbs of the same functional group, *Plantago lanceolata* and *Lotus*
219 *corniculatus*; (3) H4, composed of the two dwarf shrubs in H2, the hairgrass
220 *Deschampsia flexuosa* and the bentgrass *Agrostis stolonifera*, the latter two species
221 belonging to different functional groups, and (4) monocultures of *C. vulgaris*, *V.*
222 *myrtillus*, *A. elatius* and *H. lanatus*.

223 Biomass samples were collected three times during the drought on 30 May in
224 week 2, 27 June in week 6 and 8 Aug in week 12. The biomass data were calculated as
225 vegetative biomass in g per 0.08 m² (20 × 40 cm). The number of samples that could be
226 collected progressively decreased after week 2 due to increasing mortality.

227 2.4. Biomass and elemental analyses

228 Photosynthetic tissues were dried and pulverized in a heater for 48 h at 70 °C and
229 conserved in desiccators until analysis (<15 days). For the analyses of C and N, 0.7 mg
230 of dried, pulverized samples was weighed with a Microbalance MX5 Mettler Toledo,
231 and the concentrations of the elements were determined by combustion coupled to

232 gas chromatography (Smith and Tabatabai, 2004) with a CHNS Eurovector 3011
233 Elemental Analyzer and a Thermo Electron NA 2100 Gas Chromatograph (C.E.
234 instruments-Thermo Electron, Milan, Italy).

235 For the other elements (P, S, Ca, Mg and Fe), 0.25 g of dried, pulverized sample
236 was diluted in an acidic mixture of HNO₃ (60%) and H₂O₂ (30% w/v) and digested in a
237 microwave system (MARSXpress, CEM Corporation, Matthews, NC). The digested
238 solutions were dissolved in a 1% HNO₃ solution and then brought to final volume of 50
239 mL with ultrapure water. Blank solutions (5 mL of HNO₃ and 2 mL H₂O₂ but no sample
240 biomass) were regularly analyzed. After digestion, the concentrations of Ca, K, Mg, S, P
241 and Fe were analyzed by ICP-OES (Inductively Coupled Plasma - Optical Emission
242 Spectrometry) (Priester et al., 2011) with a spectroscopy Optima 4300DV (Perkin-Elmer
243 Inc., Wellesley, MA). We used the standard certified biomass NIST 1573a (tomato leaf)
244 to assess the accuracy of the biomass digestion and analytical procedures.

245

246 2.5. Statistical analyses

247 We conducted univariate analysis by factorial analysis of variance (ANOVA) with
248 species, community composition and time of sampling as independent variables, and
249 chemical characteristics as dependent variables for the species growing in two or more
250 communities. For the species present only in one community, we conducted ANOVAs
251 with time of sampling as an independent variable and chemical characteristics as
252 dependent variables. We used Fisher's post-hoc tests to compare the different levels
253 of the independent variable.

254 We also conducted multivariate ordination analyses by using principal
255 component analyses (PCAs) including the elemental concentrations and the N:P, N:K
256 and P:K ratios to determine: (i) whether the species growing in different communities
257 have different elemental composition and whether the changes in response to drought
258 were different depending on the community where the target species grew, by
259 analyzing species growing in different communities, and (ii) whether each of the three
260 factors studied (species, community and length of drought) had by themselves
261 significant effects on elemental composition and stoichiometry and whether species-
262 specific identity remained in all community and drought situations by analyzing the
263 data of all species and communities together.

264 We also conducted an ANOVA with the scores of the first six PC axes as
265 dependent variables and the various plant communities as fixed categorical
266 independent factors to determine the global effects of the community on elemental
267 concentrations and stoichiometries. We tested for normality and homogeneity of
268 variance of the residuals prior to the statistical analyses by examining the residuals
269 versus the expected plots and the normal qq-plots of the linear models. The data were
270 log-transformed if the required conditions were not met. The significance level was set
271 at $P<0.05$. We also correlated biomass survival (ratio of live to dead biomass) during
272 the drought with the elemental and stoichiometric PC scores to determine how
273 mortality affected stoichiometric composition. All statistical analyses were performed
274 using Statistica 6.0 (StatSoft, Inc. 1984-2001, Tulsa, OK, USA).

275 **3. Results**

276 3.1. Effects of the drought on the communities

277 Species responded to drought with a general trend of decreasing P, S and K
278 concentrations and increasing C concentrations in all four species studied in most
279 communities, and with decreasing N in three of the four studied species. Only *V.*
280 *myrtillus* N concentrations did not change in any community in response to drought
281 time increase (Table S2).

282 The PCAs indicated that overall elemental compositions generally varied in the
283 shrubs *C. vulgaris* and *V. myrtillus* in the H2 community with the advance of the
284 drought period (Figure 1a,b). Plants of different communities were separated along
285 the PC1 axis, but the overall elemental composition of both species shifted towards
286 lower concentrations of P, S and K and higher C concentrations with increasing
287 drought.

288 These two species had similar responses in foliar elemental composition in the
289 H4 community. This community also included two grasses, *D. flexuosa* and *A.*
290 *stolonifera*, whose overall elemental compositions differed from those of the two
291 shrubs along the PC1 axis ($P<0.0001$) (Figure 1c,d). The PC2 axis separated *C. vulgaris*
292 from *V. myrtillus* ($P<0.0001$), within the same functional group, and *D. flexuosa* from *A.*
293 *stolonifera* ($P<0.0001$), in different functional groups (Figure 1c,d). The responses of
294 the species to the drought were consistent with the univariate analyses (Table S2);
295 there was a trend towards lower S, and P concentrations and higher C concentrations
296 across PC2 axis with increasing drought. All four species tended to occupy a specific
297 space in the plot of the first two PCs, indicating species-specific stoichiometries despite
298 shifts in elemental compositions during the drought.

299 The PCA for the G4 community with the four herbaceous species of two
300 functional groups, *H. lanatus*, *A. elatius*, *P. lanceolata* and *L. corniculatus*, separated all
301 four species in at least one of the first three PCs. In all four species we observed an
302 overall shift in elemental composition along the PC2 axis mainly loaded by decreasing P
303 and K concentrations and increasing C concentrations as the drought progressed
304 (Figure 1e,f), results consistent with univariate analyses that further confirmed these
305 changes with increasing drought (Table S2).

306 We also found differences during the course of the drought in *P. lanceolata*
307 and *L. corniculatus* that were growing only in G4 community. In *P. lanceolata*, C and Fe
308 concentrations were higher in week 6 and decreased toward the end of the drought.
309 The K and P concentrations responded oppositely, decreasing in the middle of the
310 drought and increasing toward the end (Figure 1e,f). The leguminous forb *L.*
311 *corniculatus* decreased its N concentration throughout the course of the drought (Data
312 not shown). The Ca concentration decreased in the grass *D. flexuosa*, and the Fe
313 concentration and N:P ratio increased (Data not shown). We could not conduct the
314 analysis for *A. stolonifera* due to the absence of replicates caused by the high mortality
315 in the first sampling of the severe drought.

316 The PCA for the four species growing in monoculture indicated comparable
317 responses to those of the H4 community, where the scores in the shrubs *V. myrtillus*
318 and *C. vulgaris* generally differed from those in the two grass species *H. lanatus* and *A.*
319 *elatius* along the PC1 axis ($P<0.0001$) (Figure 1g,h). *V. myrtillus* and *C. vulgaris* were
320 separated along the PC2 axis ($P<0.0001$). *H. lanatus* and *A. elatius* were also separated
321 along PC2 axis ($P<0.0001$). The death of many individuals of *V. myrtillus* and *A. elatius*

322 did not allow an analysis for the advanced stages of the drought. The variables that
323 mainly loaded PC1 axis were the concentrations of S, K and P, which decreased, and C
324 concentrations that increased towards shrubs, and N and Ca concentrations that
325 decreased in *C. vulgaris* and *H. lanatus*, during the drought advance (Figure 1 g,h).
326 These changes were also consistent with the observed in univariate analyses (Table
327 S2).

328 Fe concentrations increased in all species, although not always significantly, as
329 the drought advanced. These increases may have been a response to the high Fe
330 content of the soil (6 mg/g) and probably to a concentration effect from the lack of
331 water during the drought.

332 The elemental compositions of the species in all communities tended to have
333 comparable shifts as the drought progressed. There were peculiar differences in these
334 shifts depending on the community, but N, P, K, Mg and S concentrations generally
335 tended to decrease and C concentrations to increase during the drought.

336 3.2. Effects of community composition on the responses of the species to the drought

337 The elemental concentrations and ratios in all species differed significantly
338 among the three sampling times during the drought (Figures 1-4, Table S2), but the
339 differences depended on the community in which the species grew. This interaction
340 between the drought and community composition was species-specific (Table S2). The
341 interaction between the duration of the drought and community composition

342 significantly affected the elemental compositions and stoichiometric ratios in three of
343 the four species that were growing in two or more different communities, showing
344 that the community had a significant influence on the shifts in elemental composition
345 during the drought. Individuals of *C. vulgaris* and *H. lanatus* in the monocultures
346 shifted their elemental concentrations and ratios mainly linearly during the drought,
347 unlike the individuals growing in more diverse communities where elemental
348 concentrations and stoichiometries change more abruptly and irregularly.
349

350 The ANOVA of PC1 scores in *C. vulgaris* showed a separation of samples over
351 the course of the drought. PC1 axis was mainly loaded by N, K, S and P, and N:P ratio,
352 showing that N, K, S and P concentrations decreased and C concentrations and N:P
353 ratios increased over the course of the drought (Figure 2a,b), consistently with the
354 results of univariate analyses (Table S2). N:K and P:K ratios peaked in the last week of
355 the drought. The shifts in elemental composition over the course of the drought,
356 though, differed among the communities containing *C. vulgaris*. The N:K and P:K ratios
357 and in general all elemental concentrations in this shrub increased in response to
358 drought more in the H4 community than in the other two communities (Figure 2a,b,
359 Table S2). The K concentration decreased in the shrub *V. myrtillus* in week 6 but
360 increased toward the end of the drought. P and S concentrations decreased, and the
361 N:P ratios increased, during the drought. This species, however, could not be studied
362 appropriately due the lack of samples from the monoculture for the final week of the
363 drought from high mortality (Figure 2c,d). Large changes in elemental composition
364 were observed between the second and sixth weeks of the drought, whereas the

365 differences between the sixth and twelfth weeks were not significant. The elemental
366 compositions in *V. myrtillus* plants generally varied among the different communities

367

368

369 during the drought as observed in the corresponding PC2 axis ($P<0.0001$). Despite of
370 these differences depending on the communities where these two shrubs grew (Figure
371 2 a-d, Table S2), the experimental drought decreased P and S concentrations in both
372 species independently of the community where they grew (Table S2).

373 The elemental compositions in the two grass species *H. lanatus* and *A. elatius*
374 also generally differed between the two communities in which they grew. The plants
375 growing in monocultures had different scores in the PCA plot of the first two PCs than
376 those growing in the G4 community with the two herb species and were separated
377 significantly along the PC2 axis ($P<0.001$ and $P<0.0001$, respectively) (Figure 3). *H.*
378 *lanatus* plants in the monoculture had lower P:K, N:K and N:P ratios and lower K
379 concentration than did those in the more diverse G4 community and consistently with
380 these results we observed in the univariate analyses lower N:K and N:P ratios in
381 monoculture plants (Table S2). Lower K concentrations were observed in *A. elatius*
382 growing in monoculture (Table S2), and consistently this variable had higher loading
383 values on PC2 axis (Figure 3c,d). Both species increased their C and Fe concentrations
384 and N:P ratios and decreased their S, P, N and K concentrations over the course of the
385 drought (Table S2).

386 The PCA for the H4 community showed a general tendency for all species to
387 shift along the PC2 axis, as a result of the overall shift towards lower concentrations of
388 N, P, Mg and S, and to higher concentrations of Fe as the drought advanced (Figure 4).

389 K concentrations and N:K and P:K ratios in *V. myrtillus* and *A. elatius*, though,
390 changed in the same direction over the course of the drought independently of the
391 community. *C. vulgaris* and *H. lanatus* showed significant interactions between
392 sampling time and community. K concentrations and N:K and P:K ratios did not change
393 in the monocultured plants but changed in those growing in H2 and H4. Moreover, the
394 distribution of the means over the PC axes reflected a minor change in the foliar
395 elemental concentrations and stoichiometries for the individuals of the four species
396 growing in monocultures. All these results suggest less stoichiometric flexibility during
397 the drought in plants growing in monocultures than in those growing in more diverse
398 communities (Figures 1-3).

399

400

401 Species were grouped by functional group in the PCA analysis (Figure 4),
402 reflecting a strong hierarchical tendency in the foliar elemental contents and
403 stoichiometries determined by a phylogenetic component. Biomass survival (the ratio
404 of live to dead biomass) during the drought was correlated with elemental and
405 stoichiometric PC scores (Figure 5), indicating a relationship, in some cases, between
406 biomass mortality and elemental composition and stoichiometry.

407 When we conducted a PCA with the data of all four studied communities
408 altogether we observed that all three studied independent factors (community,
409 species and time of drought) had effects on overall elemental composition and
410 stoichiometry (Figure S1). This further showed that there is a species-specific
411 elemental composition and an effect of drought changing the species elemental

412 composition, but species-specific identity in elemental composition and stoichiometry
413 always remains despite the level of drought.

414 3.3. Pre-treatment: recurrent droughts

415

416 The effects of recurrent droughts on elemental composition were small and not detectable in some species. We found no significant
417 differences in the elemental concentrations and ratios in the two species of dwarf shrubs (*C. vulgaris* and *V. myrtillus*) between the re-current
418 drought and the control plots in the second week of the drought (first sampling) (Table S3). The grass *H. lanatus* had higher K concentrations
419 and lower N:K ratios in response to the recurrent-drought treatment, whereas *A. elatius* had lower Mg concentrations and N:K and P:K ratios
420 (Table S4). These species-specific effects, however, disappeared during the severe drought. The foliar elemental concentrations did not differ
421 between the control and drought plots in the two other species in the grassland community (G4), the herb *P. lanceolata* and the leguminous *L.*
422 *corniculatus*, and in the grass *D. flexuosa* in the H4 community.

423 **4. Discussion**424 *4.1. Pre-exposure to drought effects in the responses of grassland and shrubland to a very severe experimental drought*

425 Increasing climatic variability within and between years and reduced reliability of seasonal climatic conditions are accepted as key
426 characteristics of the ongoing climate change (IPCC, 2013). Communities of perennial plants such as grasslands and heathlands cover a large
427 part of Europe and are of economic and ecological importance. Such ecosystems are increasingly exposed to novel conditions, leading to
428 uncertainty about their future stability and functioning. Previous treatments in the experimental site and communities of this study have
429 shown surprisingly high resilience in many ecosystemic traits such as biomass, [plant cover](#), [community species composition](#), [soil enzyme](#)
430 [activity and variability of flowering phenology](#) (Kreyling et al., 2008a; Jentsch et al., 2011). Nevertheless, significant impacts were found on the
431 allocation of carbon in the soil (Mirzae et al., 2008) and on the functioning of plants within their life cycles (Jentsch et al., 2011). Pre-exposure
432 to drought enhanced physiological processes and even led to memory effects (Walter et al., 2011, 2013). In the present study, however, the
433 previous recurrent droughts (pre-treatment) had no significant effect on the shifts in elemental composition during the severe drought.

434

435 *4.2. Species-specific differences*

436 Despite the dependence of the responses to the drought on community composition, the shifts in elemental compositions during the extended
437 severe drought tended to have common trends among the species. The drought had general effects on the foliar compositions of the species: a
438 decrease in the N, P, K concentrations and an increase in C concentration and in C:nutrient and N:P ratios. As expected, the nutritional status of
439 photosynthetic tissues became increasingly impoverished with the advancement of the drought, coinciding with an increase in mortality and a
440 deterioration of the health status of plants. All species had a better nutritional status at the time of the first sampling two weeks after the
441 onset of the drought than at the final sampling 12 weeks after the onset. This finding can be appreciated in the PCA plots, where all individuals
442 of the first sampling, regardless of community, tended to occupy a PCA space that reflected higher concentrations of macronutrients,
443 irrespective of the dissimilarities among the different communities. The shifts in the foliar stoichiometries of the species, however, became less
444 predictable in the more diverse communities as the drought advanced.

445 Some degree of species-specific response was evident in this situation of extreme stress. *C. vulgaris* was more drought tolerant than *V.*
446 *myrtillus*. Its elemental concentrations suggested a large sclerophyllous development in all communities to confront water stress. In contrast,
447 all *V. myrtillus* plants died in the monoculture before the final week of drought, suggesting that cohabitation with other species favored
448 survival, reinforcing the significant role of diversity. The large variation in the foliar stoichiometry of *H. lanatus* during the drought indicated a
449 high capacity to modify elemental concentrations (high stoichiometric flexibility). The high K concentrations in the final week were consistent

450 with the importance of this element for the conservation of water. The other grass species, *A. elatius*, had the highest C and N concentrations
451 by the end of the drought, indicating that this species responded differently to the water stress and was better adapted to drought conditions.
452 K accordingly plays a fundamental role in terrestrial environments under conditions of water stress. Water availability is the main limiting
453 factor in terrestrial ecosystems, so analyses of K should be included in ecological stoichiometric studies (Rivas-Ubach et al., 2012; Sardans et al.,
454 2012c). Our results are consistent with the postulate by Yu et al. (2011) that plants with lower N:P ratios, the grass species in our study (7-8.5 in
455 *A. elatius*, 7.5-9.2 in *D. flexuosa* and 7-7.5 in *H. lanatus*), should have more flexible stoichiometries than plants with higher N:P ratios, the shrub
456 species (11.4-12.7 in *C. vulgaris* and 10.7-12.6 in *V. myrtillus*).

457

458 *4.3. Effects of community composition*

459 The changes in elemental compositions during the drought among the various species growing in the same communities were generally quite
460 similar, but the grasses were usually more plastic than the shrubs and herbs. The responses to the drought, however, varied among the
461 different communities. For example, P concentrations in all communities and species decreased in response to the drought, but the changes in
462 K concentration depended on the community. K concentrations did not change in H4 but decreased in H2 over the course of the drought. *C.*

463 *vulgaris* and *V. myrtillus*, which were present in these two communities, thus had different elemental changes depending on the community
464 composition in which they grew. These results provided strong evidence for higher similarities in the response to the drought in species
465 growing in the same community compared to the same species growing in different communities, showing that stoichiometric shifts in
466 response to drought are more-community dependent than species-dependent.

467 The interaction between the time during the drought and community composition was significant for several elemental concentrations
468 in *C. vulgaris* and *H. lanatus* and for one ratio in *A. elatius*. These interactions supported our hypothesis that community composition influences
469 the changes in plant stoichiometry under extreme water stress. The species growing in monocultures tended to have a less plastic response
470 and a more consistent pattern of change in elemental concentrations and stoichiometries during the drought than did the same species
471 growing in communities. The elemental compositions in species growing in a richer community, however, tended to change more abruptly and
472 erratically, probably by adjusting their stoichiometries to the variable demands and requirements under a more variable competitive
473 environment. Some studies have observed that plants in highly diverse communities are able to take up more nutrients than are plants in less
474 diverse communities (Massey et al., 2013). Wisheu et al. (2000) reported that nutrient-poor Mediterranean ecosystems have a high diversity of
475 plants due to the large number of specialists able to take up nutrients in different ways and at different soil depths. Under a scenario of global
476 change with more sudden intense and severe droughts and an increasing probability of nutrient imbalances (Peñuelas et al., 2012; 2013),

477 communities with more species but with fewer individuals of each species could thus be expected to be more successful, at least over short
478 time scales, because these communities would be better able to respond quickly by using the resources more efficiently. A recent review has
479 reported that plant species diversity is consistently higher in infertile soils, an association linked to functional diversity for nutritional strategies
480 (Lambers et al., 2011).

481 Community composition and diversity of functional groups had significant influences on tissue dieback in the communities in previous
482 treatments (Kreyling et al., 2008b; Backhaus et al., 2014). We have also observed that the dieback of biomass in some communities and species
483 was closely associated with the changes in elemental stoichiometry over the course of the drought, whereas this relationship was not as clear
484 in the same species growing in other communities. Furthermore, soil microbial communities and their contributions to nutrient cycling and
485 uptake are modified by the presence or absence of particular plant species rather than by species richness (Bremer et al., 2009).

486 The better status and contents of nutrients in plants growing in communities with more species suggests that interspecific competition
487 can have some advantages over intraspecific competition in times of drought; the more variable demands of soil resources can lead to different
488 abiotic requirements, such as water, nutrients and space that can be freely used by the diverse species that compose the community. A higher
489 stoichiometric flexibility is more likely in a more diverse community, which is advantageous for responding to environmental changes. Species
490 intolerant to stress could take advantage of this higher stoichiometric plasticity to avoid competition with stress-tolerant neighbors that

491 respond to stress by increasing sclerophyll and thus lowering their demands for resources, thereby increasing the availability of resources to
492 the intolerant species. Each species exhibited specific elemental stoichiometries, with elemental compositions more similar between herbs and
493 grasses than between these species and the shrubs. The differences among species were independent of the composition of the community in
494 which the plants grew and were consistent with those in a previous study conducted before the severe drought (Urbina et al., 2014). **Although,**
495 **the study did not include data of roots, leaves (or photosynthetic tissues) constitute the crucial organ when plant responds to changes in its**
496 **functions. In this context, we can not define the plant global elemental composition and stoichiometry responses but by knowing the values**
497 **and changes of foliar biomass stoichiometry we can define the "species biogeochemical niche". It represents the adequate foliar elemental**
498 **composition and stoichiometry for the optimal foliar function for a determined species in its environmental circumstances.**

499

500 **5. Final remarks and conclusions**

501 The responses to a very severe drought were more similar in different species growing in the same community than in the same species
502 growing in different communities, indicating that stoichiometric shifts in response to drought were more community-dependent than species-

503 dependent. The degree (range) of the change, however, varied among species, with grasses being more plastic than shrubs or herbs. Previous
504 recurrent droughts (pre-treatments) had no significant effects on the shifts in elemental composition during the severe drought.

505 This study provides evidence of the suitability of studying the differences and shifts in plant elemental composition to improve our
506 understanding of the processes underlying the responses of plants to different biotic and abiotic environmental circumstances. Species have
507 particular foliar elemental compositions and stoichiometries, depending in part on their ecological lifestyle, but they also have some degree of
508 flexibility to adapt to changes in abiotic conditions (e.g. drought) and biotic conditions (e.g. community composition). Our results are consistent
509 with the biogeochemical niche hypothesis (Peñuelas et al., 2008; 2010). Complex communities have a higher capacity to respond to drought
510 relative to stands dominated by only one species. Species-specific abilities to exploit the available resources may explain this phenomenon.

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662

663 Figure captions:

664 Figure 1. PCA plots of the mean \pm 95% confidence interval of the scores of first two PCs for the species during the course of the severe drought in the H2 (a),
 665 H4 (c) and G4 (e) communities and the monocultures (g), and the corresponding variable distributions in H2 (b), H4 (d), G4 (f) and the monocultures (h). The
 666 ellipses represent the 95% confidence regions around the point representing the scores mean of PC1 and PC2 axes for each species. Percentages

667 within brackets in the X and Y axes legends indicate the percentage of variance explained by the corresponding PC. Codes: number of weeks since the
 668 beginning of drought (two weeks = 2, six weeks = 6, twelve weeks = 12). Species (*C. vulgaris* = **C**, *V. myrtillus* = **V**, *A. stolonifera* = **Ag**, *D. flexuosa* = **D**, *H.*
 669 *lanatus* = **H**, *A. elatius* = **A**, *L. corniculatus* = **L**, *P. lanceolata* = **P**).
 670

671

672 Figure 2. PCA plots of the mean \pm 95% confidence interval of the scores of the first two PCs for *Calluna vulgaris* (a) and *Vaccinium myrtillus* (c) growing in
 673 different communities (diversity levels) during the course of the severe drought, and the corresponding variable distributions (b) and (d), respectively. The
 674 ellipses represent the 95% confidence regions around the point representing the scores mean of PC1 and PC2 axes for each species. Percentages
 675 within brackets in the X and Y axes legends indicate the percentage of variance explained by the corresponding PC. Codes: number of weeks since the
 676 beginning of drought (two weeks = 2, six weeks = 6, twelve weeks = 12). Communities (**H2 community** = **H2**, **H4 community** = **H4**, Monoculture = **M**).
 677

677 Figure 3. PCA plots of the mean \pm 95% confidence interval of the scores of the first two PCs for *Holcus lanatus* (a) and *Arrhenatherum elatius* (c) growing in
 678 different communities (diversity levels) during the course of the severe drought, and the corresponding variable distributions (b) and (d), respectively. The
 679 ellipses represent the 95% confidence regions around the point representing the scores mean of PC1 and PC2 axes for each species. Percentages
 680 within brackets in the X and Y axes legends indicate the percentage of variance explained by the corresponding PC. Codes: number of weeks since the
 681 beginning of drought (two weeks = 2, six weeks = 6, twelve weeks = 12). Communities (**G4 community** = **G4**, Monoculture = **M**).
 682

682 Figure 4. PCA plots of the mean \pm S.E. scores of the first two PCs for the two shrub and two grass species (a) growing in the H4 community during the course
 683 of the severe drought, and the corresponding variable distributions (b). Arrows on the X axis of the mean PC scores indicate the averages of each species
 684 group (shrubs and grasses) when the differences between them were significant ($P < 0.05$). When there are more than two species, different letters indicate
 685 statistical significances ($P < 0.05$) of the differences among them. Percentages within brackets in the X and Y axes legends indicate the percentage of variance
 686 explained by the corresponding PC.
 687

687 Figure 5. Relationships between the PC1 scores in Figure 2 and biomass survival (BS; ratio of live to dead biomass) during the course of the severe drought
 688 for *Calluna vulgaris* growing in the H2 community (a) and for *Arrhenatherum elatius* growing in the G4 community (b). Percentages within brackets in the X
 689 and Y axes legends indicate the of variance explained by the corresponding PC.
 690

Table 1. Species compositions of the communities.

Community	Species
H2	<i>Calluna vulgaris, Vaccinium myrtillus</i>
G4	<i>Arrhenatherum elatius, Holcus lanatus, Plantago lanceolata, Lotus corniculatus</i>
H4	<i>Calluna vulgaris, Vaccinium myrtillus, Deschampsia flexuosa, Agrostis stolonifera</i>
Monocultures	<i>Calluna vulgaris</i>
	<i>Vaccinium myrtillus</i>
	<i>Arrhenatherum elatius</i>
	<i>Holcus lanatus</i>

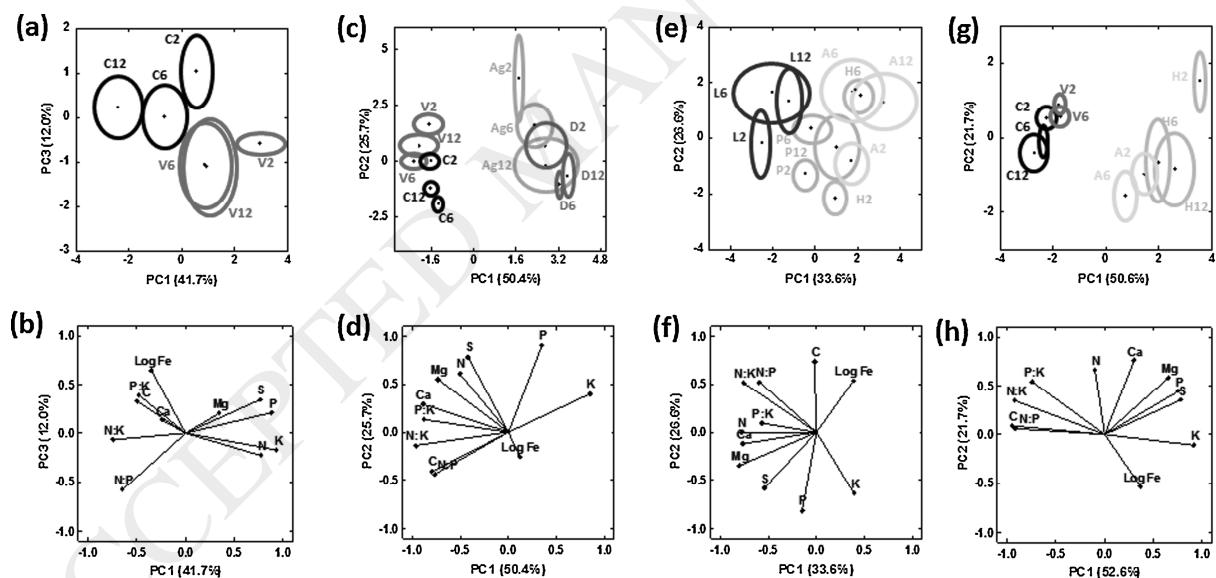


Fig. 1

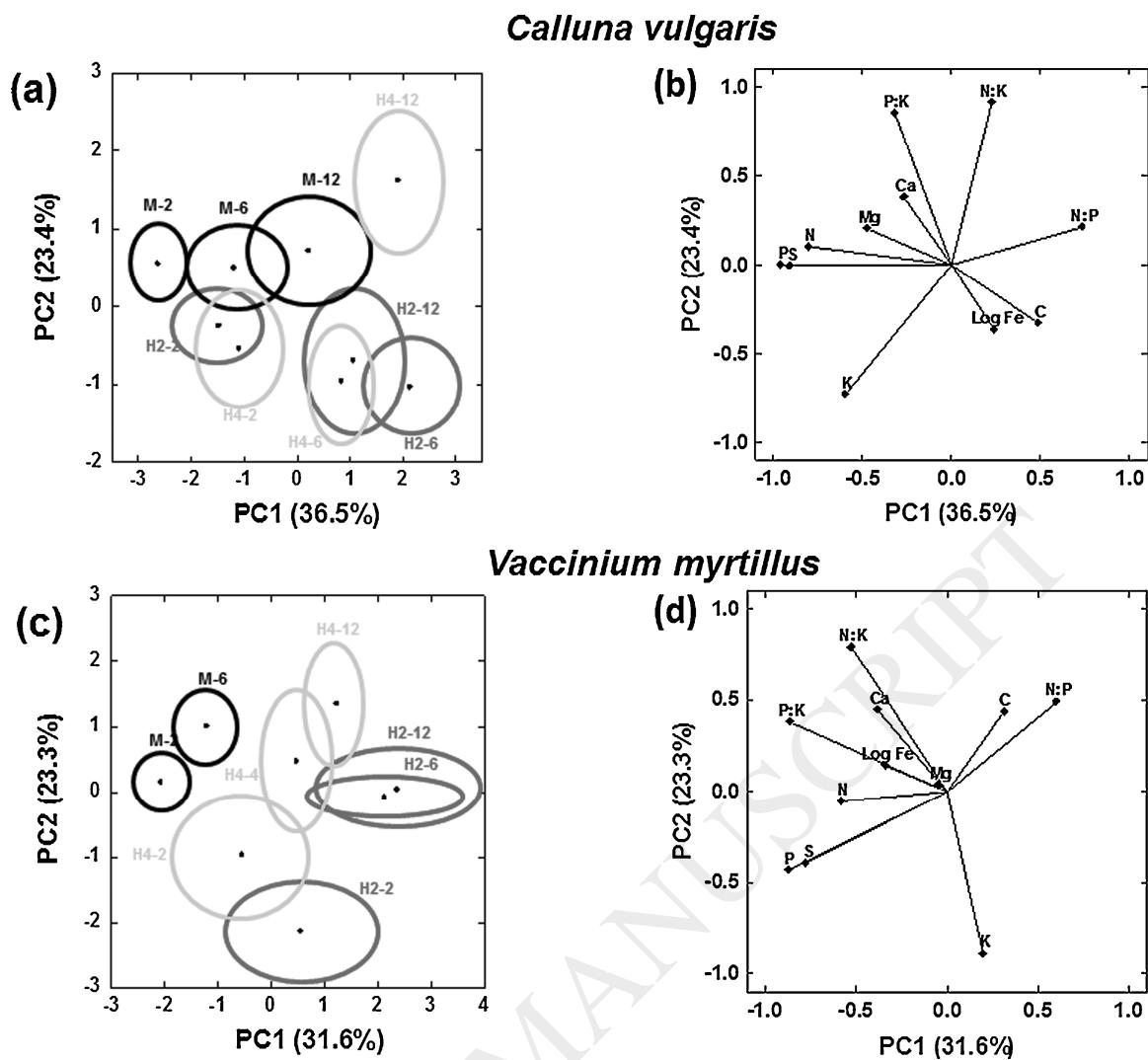


Fig. 2

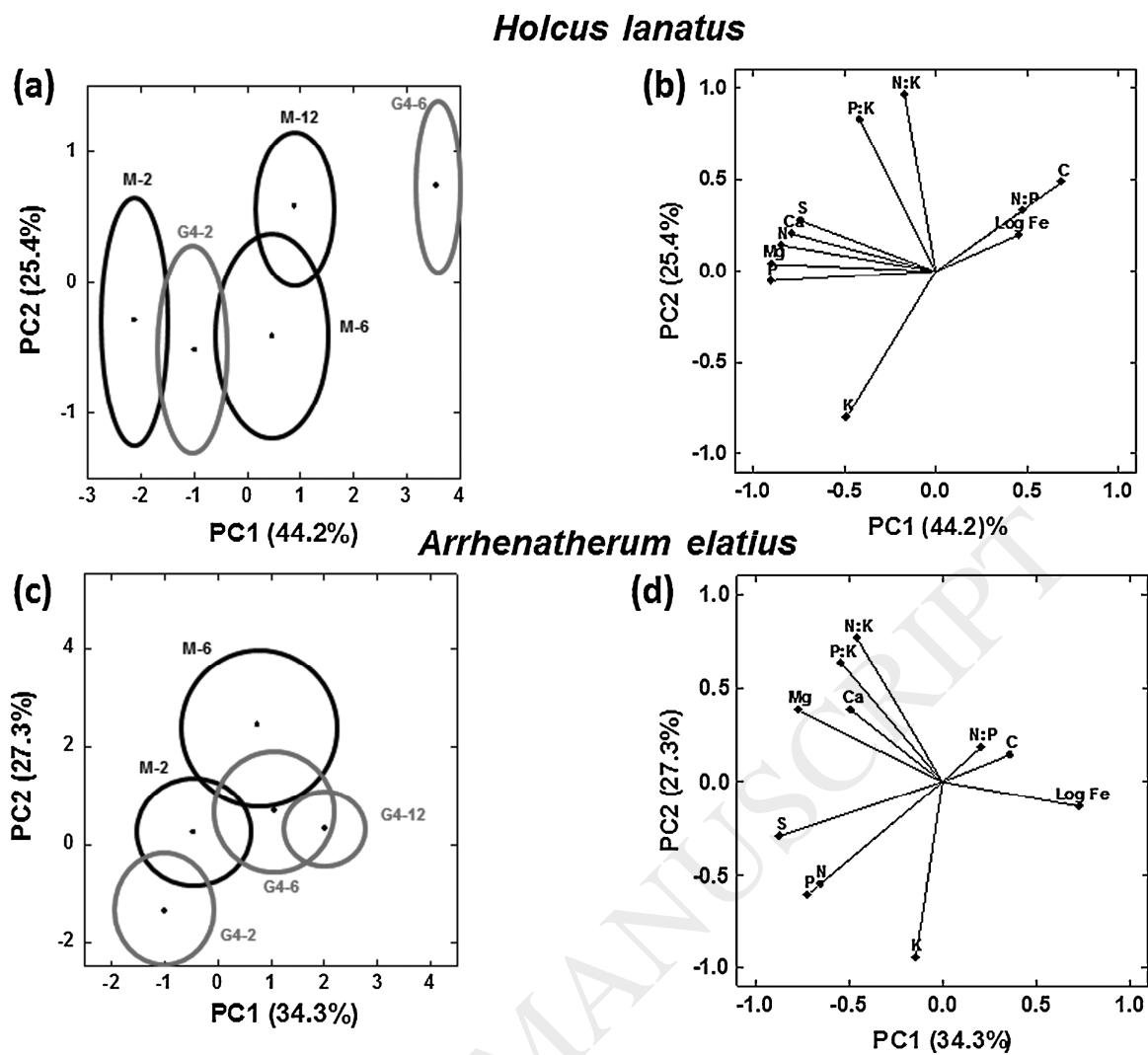


Fig. 3

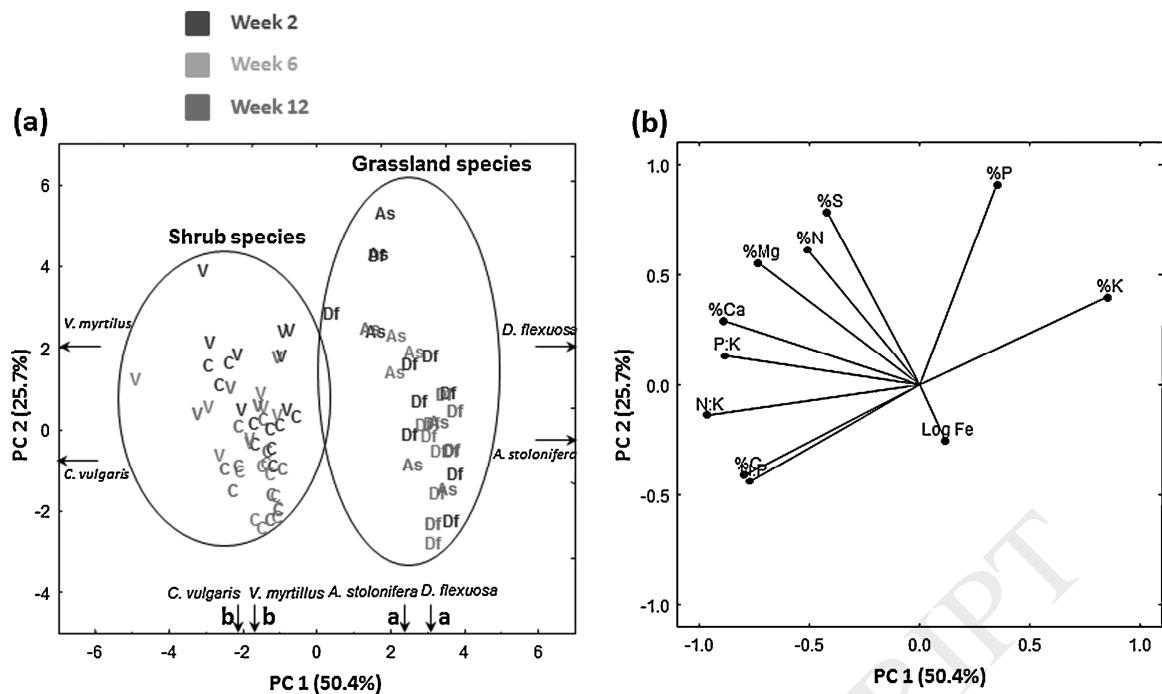


Fig. 4

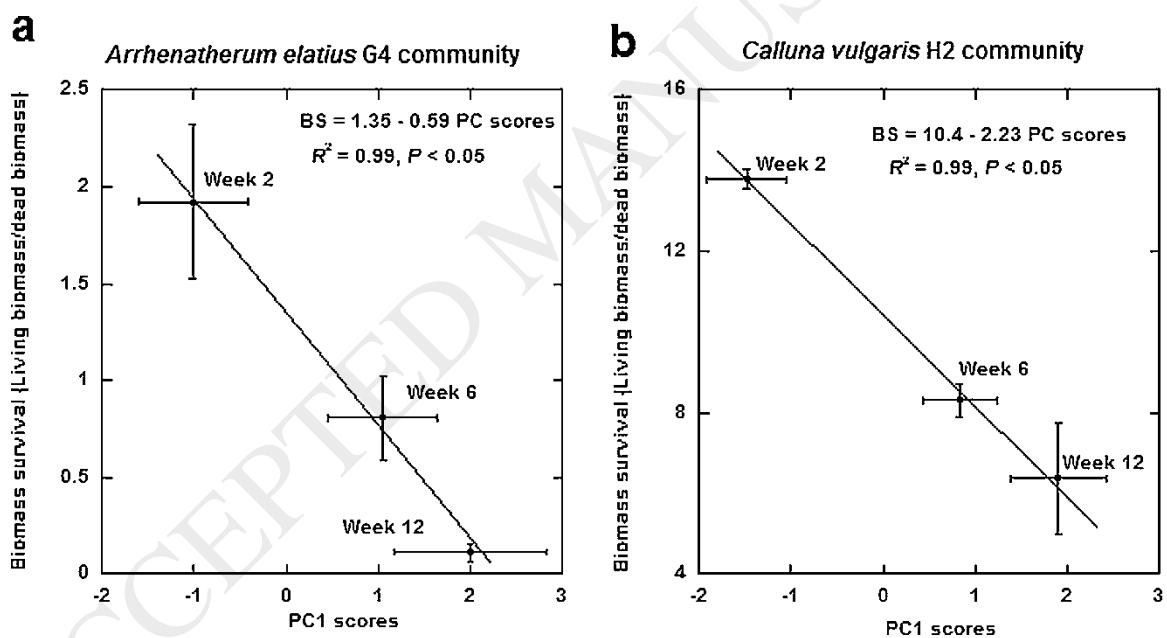


Fig. 5