

This is the accepted version of the following article:

Villagrasa E., Ballesteros B., Obiol A., Millach L., Esteve I., Solé A. . Multi-approach analysis to assess the chromium(III) immobilization by *Ochrobactrum anthropi* DE2010. *Chemosphere*, (2020). 238. 124663: - .
10.1016/j.chemosphere.2019.124663,

which has been published in final form at
<https://dx.doi.org/10.1016/j.chemosphere.2019.124663> ©
<https://dx.doi.org/10.1016/j.chemosphere.2019.124663>. This manuscript version is made available under the CC-BY-NC-ND 4.0 license
<http://creativecommons.org/licenses/by-nc-nd/4.0/>

1 **Title: Multi-approach analysis to assess the chromium(III) immobilization**
2 **by *Ochrobactrum anthropi* DE2010**

3

4 **Abstract**

5 *Ochrobactrum anthropi* DE2010 is a microorganism isolated from Ebro
6 Delta microbial mats and able to resist high doses of chromium(III) due to its
7 capacity to tolerate, absorb and accumulate this metal. The effect of this
8 pollutant on *O. anthropi* DE2010 has been studied assessing changes in
9 viability and biomass, sorption yields and removal efficiencies. Furthermore,
10 and for the first time, its capacity for immobilizing Cr(III) from culture media was
11 tested by a combination of High Angle Annular Dark Field (HAADF) Scanning
12 Transmission Electron Microscopy (STEM) imaging coupled to Energy
13 Dispersive X-ray spectroscopy (EDX).

14 The results showed that *O. anthropi* DE2010 was grown optimally at 0-2
15 mM Cr(III). On the other hand, from 2 to 10 mM Cr(III) microbial plate counts,
16 growth rates, cell viability, and biomass decreased while extracellular polymeric
17 substances (EPS) production increases. Furthermore, this bacterium had a
18 great ability to remove Cr(III) at 10 mM ($q= 950.00 \text{ mg g}^{-1}$) immobilizing it mostly
19 in bright polyphosphate inclusions and secondarily on the cellular surface at the
20 EPS level. Based on these results, *O. anthropi* DE2010 could be considered as
21 a potential agent for bioremediation in Cr(III) contaminated environments.

22 **Keywords:** *Ochrobactrum anthropi* DE2010; CLSM; ICP-OES; HAADF-STEM
23 EDX; polyphosphate inclusions

24 **1. Introduction**

25 The increase in industrial and anthropogenic activities has largely
26 contributed towards the introduction of heavy metals and other pollutants into
27 the environment (Gupta et al., 2014; Bhattacharya et al., 2015). Pollution by
28 heavy metals has become a serious threat to the environment and public
29 health, since, in general, they are highly toxic and accumulate throughout the
30 food chain (Feng et al., 2012; Xu et al., 2014).

31 Chromium is one of the earth crust's most abundant elements (Mohana
32 and Pittman, 2006) and it is a redox active 3d transition metal with different
33 oxidation states ranging from 2+ to 6+ (Greenwood and Earnshaw, 1998;
34 Srivastava and Thakur, 2007). The trivalent (Cr(III)) and hexavalent (Cr(VI))
35 forms, from industrial by-products or by being part of phosphate fertilizers used
36 in agriculture (Nziguheba and Smolders, 2008; Manahan, 2009) among other
37 chromium sources, are stable in the majority of terrestrial surface and aquatic
38 environments (Kimbrough et al., 1999; Ihsanullah et al., 2016). Cr(III) is
39 considered to be an essential element (micronutrient) with known cellular
40 biological functions (Balk et al., 2007; Evert et al., 2013). However, with long-
41 term exposure it can produce toxic effects in the cells, but at a much lower
42 degree than Cr(VI), which is more toxic and unhealthy for living beings and the
43 environment (Shanker et al., 2005; Francisco et al., 2011; Ihsanullah et al.,
44 2016). Even so, Cr(III) is involved in some human diseases such as: structural
45 perturbation in erythrocyte membrane and cancer (Kusiak et al., 1993;
46 Suwalsky et al., 2008; Figgitt et al., 2010).

47 In the last decade, our research group has isolated different
48 heterotrophic and phototrophic (cyanobacteria and algae) microorganisms from
49 the Ebro Delta microbial mats (Tarragona, Spain). One of these

50 microorganisms, *Ochrobactrum anthropi* DE2010, was isolated from the
51 microalgae *Scenedesmus* sp. DE2009 consortium from the same ecosystems,
52 and recently was characterized and identified by means of microbiological,
53 biochemical and molecular methods. Furthermore, the analysis of the effect of
54 deprived nitrogen source conditions on this heterotrophic bacterium
55 demonstrate that it is able to overcome this limiting condition through a still
56 unknown *nifH*-independent mechanism, although there were changes in shape,
57 size, and abundance of pleomorphic cells (Villagrasa et al., 2019).

58 The aims of this work are: (i) to evaluate the cytotoxic effect of chromium
59 and the capacity of *Ochrobactrum anthropi* DE2010, growing in axenic cultures,
60 to remove it; (ii) to determine changes in total biomass and cellular viability and
61 in composition and production of EPS in cultures exposed to this metal and (iii)
62 to determine the Cr(III) uptake efficiency of *Ochrobactrum anthropi* DE2010 and
63 its ability to capture chromium(III) extra and/or intracellularly. With this in mind,
64 an analytical multi-approach combining classical microbial plate counts, growth
65 methods, optical microscope techniques, and biochemical and chemical
66 analysis besides high resolution microscopy techniques have been applied.

67 **2. Materials and methods**

68 2.1 Microorganism, chromium(III) stock solutions and culture conditions

69 *O. anthropi* DE2010 (accession number DDBJ/ENA/GenBank,
70 KY575285) isolated from a *Scenedesmus* sp. DE2009 consortium from Ebro
71 Delta microbial mats (Villagrasa et al., 2019) was grown at 27 °C in Luria–
72 Bertani (LB) agar medium, containing tryptone (10.0 g L⁻¹), yeast extract (5.0 g

73 L⁻¹), NaCl (10.0 g L⁻¹) and agar (15.0 g L⁻¹), and preserved in Cryoinstant® vials
74 (Thermo Fisher Scientific) at -80 °C.

75

76 Chromium(III) stock solution was prepared as Cr(NO₃)₃· 9H₂O salt
77 (Sigma-Aldrich, Bellefonte, PA, USA). The 2,600 mg L⁻¹ (50 mM) Cr(III) stock
78 solution was made by dissolving the exact quantities of Cr(NO₃)₃ in double
79 deionized water and sterilized by filtration through a 0.2 µm filter (Millipore,
80 Merck Millipore). The Cr(III) working concentrations of 0 (control experiment),
81 0.5, 2, 5, 7, and 10 mM (equivalent to 0, 26, 104, 260, 364, and 520 mg L⁻¹,
82 respectively) were obtained by serial dilution. The Cr(III) stock and serial
83 working solutions were prepared just before use and its pH were adjusted at
84 6.5.

85

86 Unpolluted (0 mM) and polluted (0.5, 2, 5, 7 and 10 mM) cultures were
87 prepared at the same conditions in the following manner: 2 mL of a 24 h culture
88 of *O. anthropi* DE2010 grown in LB (OD₆₀₀ between 1.4-1.6, approximately 10¹⁰
89 cfu mL⁻¹) was inoculated into 18 mL of LB liquid medium with the different
90 tested Cr(III) concentrations. These cultures were used for all experiments and
91 were incubated in an orbital shaker (Infors HT, Ecotron) (150 rpm) at 27 °C for
92 24 h. Triplicate cultures were grown for each heavy metal concentration.

93

94 2.2 Bacterial plate counts and growth curves

95 To determine the concentration of bacterial cells, 0.1 mL from 24 h
96 cultures grown at different Cr(III) concentrations were spread onto LB agar
97 plates also supplemented with the same final chromium working concentrations:

98 0, 0.5, 2, 5, 7, and 10 mM. In addition, in order to evaluate the number of
99 dormant *O. anthropi* DE2010 cells, 0.1 mL from each 24 h culture polluted with
100 Cr(III) was spread onto new LB agar plates without Cr(III). The plates were
101 incubated at 27 °C for 2 d in darkness. The viability counts were expressed in
102 cfu/mL, and 8 replicates were performed for each experiment. The dormant cell
103 counts were determined using the following formula (Sachidanandham et al.,
104 2009):

$$105 \quad DC = [100 - (\frac{CC}{CC + RC} \times 100)]$$

106 where DC is the dormant cells relative ratio; CC is the cultivable cells grown
107 with Cr (III) count (cfu mL⁻¹) and RC is the resurrected cells count (cfu mL⁻¹) and
108 corresponds to the difference between cfu mL⁻¹ obtained without and with Cr
109 (III).

110
111 For growth curve measurements, *Ochrobactrum anthropi* DE2010 culture
112 was grown in LB medium at 27 °C for 24 h in the dark. After this incubation
113 time, a 96-well microplate was prepared by inoculating 20 µL of the above-
114 mentioned over-night culture in microplate columns 2, 4, 6, 8, 10, and 12 filling
115 all the wells. Working Cr(III) concentrations 0, 0.5, 2, 5, 7, and 10 mM at a final
116 volume of 200 µL were used. Blank media samples were considerate by adding
117 bacteria-free LB medium uncontaminated and contaminated with Cr(III) using
118 the same working concentrations in microplate columns 1, 3, 5, 7, 9, and 11,
119 respectively. OD (λ = 600 nm) was measured every 30 min during 24 h with a
120 microtiter plate photometer (Thermo Scientific Variokan® Flash, Waltham,
121 USA). Before each measurement, microtiter plate was shaken for 5 s at 120
122 rpm. Eight replicates were analyzed for each heavy metal concentration.

123

124 2.3 Cell viability and biomass analysis by Confocal Laser Microscopy (CLSM)

125 Biomass and viability of *O. anthropi* DE2010 cells at different Cr(III)

126 concentrations were determined according to the fluorochromes-confocal laser

127 scanning-image analysis method (FLU-CLSM-IA) described by Puyen et al.

128 (2012a) with some modifications introduced for this particular study. This

129 method combined the use of specific fluorochromes, CLSM, and image analysis

130 using the ImageJ/FIJI v.1.46r software (Schneider et al., 2012).

131

132 For this experiment, 1 mL from each 24 h culture contaminated with a

133 different Cr(III) working concentration was centrifuged at 2000xg for 10 min at 4

134 °C (Eppendorf 5804R refrigerated centrifuge). Then, the obtained pellets were

135 stained with a mixture of 100 µL of two fluorochromes, SYTOX® Green (5 µM)

136 and Hoechst 33342 (64.9 µM) for 30 min at room temperature in the dark.

137 Subsequently, 25 µL of these stained pellets were smeared onto Polylysine®

138 slides (ThermoFisher Scientific, USA), producing a sticky surface and

139 preventing cell movement, mounted in BacLight mounting oil (ThermoFisher

140 Scientific, USA) and immediately covered and sealed with coverslips (24x24

141 mm). Stained cells were observed with a confocal microscope (Leica TCS SP5;

142 Leica Microsystems CMS GmbH, Mannheim, Germany) using a HCX PL APO

143 lambda blue 63.0x1.40 OIL UV at 2 zoom. In this study, xyz CLSM images were

144 taken in a sequential scan in two channels to distinguish the fluorescence

145 emitted by Hoechst 33342 (414-485 nm) and SYTOX Green (515-580 nm),

146 respectively. In order to differentiate between living and dead cells, they were

147 stained with Hoechst 33342 or SYTOX Green respectively, blue (live cells) and

148 green (dead cells) pseudo-colors were used. In each bacterial culture, once the
149 20 blue and green CLSM images (1024x1024 pixels) were acquired, the
150 corresponding CLSM binary images (black/white) were generated by means of
151 the ImageJ/FIJI v.1.46r program. In addition, Voxel counter plugin of this
152 software was applied to assess the cellular viability and biomass at each Cr(III)
153 concentration (Puyen et al., 2012a).

154

155 2.4 Extraction and Biochemical analysis of the total EPS

156 Total EPS of *O. anthropi* DE2010 from all analyzed cultures were
157 extracted using the physico-chemical method described by Adav and Lee
158 (2008), with some modifications. For this analysis, 20 mL from each
159 contaminated (0.5, 2, 5, 7, and 10 mM Cr(III)), and non-contaminated (control)
160 culture were centrifuged at 2000xg for 10 min at 4 °C (Eppendorf 5804R
161 refrigerated centrifuge), and pellets were resuspended in 10 mL of sterile
162 double deionized water. Cellular suspensions were mixed with 60 µL of ultra-
163 pure formamide (ThermoFisher Scientific, USA) and incubated on ice for 60
164 min. Then, 4 mL of 1N NaOH was added and the resulting mixtures were
165 incubated for 40 min on ice, and afterward were subjected to ultrasound at 120
166 W for 5 min on ice in an ultrasonic bath (Sonorex, Bandelin). All sonicated
167 suspensions were centrifuged at 10000xg for 20 min at 4 °C. The supernatants
168 obtained were filtered, to avoid the presence of cells, through 0.22 µm pore size
169 filters (Millipore, Merck Millipore) to collect the total EPS. Aliquots of 5 mL of
170 each of the EPS extracts were stored at -20 °C until further analysis.
171 Biochemical composition of each EPS extract was determined using various
172 colorimetric methods as follows: total carbohydrate content with the phenol–

173 sulphuric acid method (Dubois et al., 1956), using glucose (1 mg mL⁻¹, Merck)
174 as standard; protein content by the Bradford method (Bradford, 1976), with
175 bovine serum albumin (2 mg mL⁻¹, Pierce) as standard, and the uronic acids
176 content by the *m*-hydroxyphenyl method (Kintner and Van Buren, 1982), using
177 galacturonic acid (1 mg mL⁻¹, Fluka) as standard. The cell lysis caused by EPS
178 extraction was evaluated by quantification of the DNA present in the extracts by
179 the diphenylamine colorimetric method (Burton, 1956) using salmon sperm DNA
180 (1 mg mL⁻¹, Sigma) as standard. Low content of DNA (1-1.2%) indicates that
181 the EPS extracted are not contaminated by intracellular substances (Liu and
182 Fang, 2002). A total of 15 replicates for each EPS extract were analyzed, with
183 the aid of a Beckman Coulter DU[®] 730 spectrophotometer (Beckman Coulter,
184 Harbor Blvd., Fullerton CA, USA). All calculations were expressed in ppm (parts
185 per million) corresponding to 1 microgram of EPS component (protein, total
186 carbohydrates or uronic acid) per milliliter of total EPS extract (µg mL⁻¹).
187 On the other hand, the production of EPS was expressed as a percentage of
188 the total concentration of EPS in relation to cell concentration.

189

190 2.5 Inductively Coupled Plasma Optical Emission Spectrometer (ICP-OES)

191 Cr (III) removal capacity of *O. anthropi* DE2010 was analyzed by ICP-
192 OES and the specific metal removal (*q*), expressed as mg of chromium/g of dry
193 weight, was calculated using the following formula (Chakravarty and Banerjee,
194 2012):

$$195 \quad q = \frac{V(C_i - C_f)}{m}$$

196 where V is the sample volume (L); C_i and C_f are the initial and final metal
197 concentrations (mg L⁻¹), and m is the dry culture biomass (g).

198

199 After 24 h, *O. anthropi* DE2010 cultures growing at each tested
200 concentration were centrifuged at 2000xg for 10 min at 4 °C (Eppendorf 5804R
201 refrigerated centrifuge). Samples from the resulting supernatants and liquid
202 samples from the initial Cr(III) working concentrations (without microorganism)
203 were analyzed by ICP-OES. Pellets were resuspended in 10 mL of sterile
204 double deionized water and the corresponding EPS extracts were obtained
205 using the same method as that described above (Adav and Lee, 2008). These
206 extracts and samples from the same cellular suspensions but without EPS were
207 also analyzed by ICP-OES, in order to detect the presence of chromium
208 immobilized into the cells. Before chemical assays, all samples were digested
209 with HNO₃ in a 600 W microwave digestion system to remove all organic
210 matter. Chromium analysis at 267.716 nm was carried out on the resultant
211 digestions using an ICP-OES spectrometer Optima 4300DV (Perkin Elmer LLC,
212 761 Main Avenue, Norwalk, USA), in quadruplicate assays.

213

214 2.6 Energy dispersive X-ray spectroscopy (EDX) analysis coupled to Scanning
215 transmission electron microscope (STEM) imaging

216 Metal sorption capability of *O. anthropi* DE2010 was confirmed by STEM
217 EDX. Samples were prepared following the TEM-EDX protocol described by
218 Maldonado et al. (2010). To this end, a sample from each 24 h culture
219 contaminated with a Cr(III) working concentration was centrifuged at 4000xg for
220 5 min at 4 °C (Eppendorf 5804R refrigerated centrifuge). The resulting pellets
221 were included in soft agar and subsequently fixed in 2.5% glutaraldehyde
222 diluted in Millonig phosphate buffer (Millonig, 1961) for 2 h and washed four

223 times (15 min) in the same buffer. Then samples were post-fixed in 1% OsO₄ at
224 4 °C for 2 h and washed again in the same buffer. They were then dehydrated
225 in a graded series of acetone (50, 70, 90, 95, and 100%), and embedded in
226 Spurr's resin. Semi-fine sections (200 nm thick) were mounted on a titanium
227 grid covered with a thin carbon layer without additional staining (lead citrate and
228 uranyl acetate). High angle annular dark field scanning transmission electron
229 microscopy (HAADF-STEM) images and STEM EDX profiles were acquired
230 using a FEI Tecnai G2 F20 microscope operated at 200 kV and equipped with
231 an EDAX super ultra-thin window (SUTW) X-ray detector (FEI, Hillsboro,
232 Oregon, USA).

233

234 2.7 Data analysis

235 Statistical analysis was carried out by one-way analysis of variance
236 (ANOVA) and Tukey multiple comparison *post-hoc* test. Significant differences
237 in ANOVA and Tukey's test were accepted at $p \leq 0.05$. The analyses were
238 performed using SPSS software (version 20.0 for Windows 7).

239 3. Results and discussion

240 In this study, several analytical methods were applied to batch cultures in
241 order to evaluate the cytotoxic effect of Cr(III) on *O. anthropi* DE2010, and the
242 capacity of this bacterium to remove it.

243

244 3.1 Changes in plate counts, cell viability and biomass

245 The changes in plate counts, growth curves, cell viability, and biomass in
246 the presence of chromium are shown in Figure 1 (Log₁₀ cfu mL⁻¹ of total and

247 cultivable cells, and growth curves), and Figure 2 (FLU-CLSM-IA imaging, live
248 and dead cells%, mg C (cm³)⁻¹, and DC%). The effect of Cr(III) on *O. anthropi*
249 DE2010 varied significantly depending on the concentration, although is more
250 evident at the highest metal concentrations. This is evidenced by a reduction in
251 the number of cfu mL⁻¹, cell viability and biomass, and an increase in DC% and
252 duplication time.

253

254 The maximum values of the four biological parameters were obtained at
255 0 mM Cr(III) (9.26x10⁹ cfu mL⁻¹ of cultivable cells, 4.02 h duplication time,
256 99.82% of live cells and 46.76 mg C (cm³)⁻¹ of biomass) and the minimum were
257 recovered out for 7 mM (<10 cfu mL⁻¹ of cultivable cells, 21 h duplication time,
258 62.72% of live cells and 3.30 mg C (cm³)⁻¹ of biomass) (Fig 1, and Fig 2).

259 Statistically significant differences were obtained, ranging from 0 mM (control)
260 to 7 mM Cr(III) for viability (F= 213.434) (*p* <0.05) and biomass (F= 212.253) (*p*
261 <0.05) results. Using the Tukey multiple comparisons *post-hoc* test, statistically
262 significant differences (*p* <0.05) were observed among all the metal
263 concentrations tested for cell viability and biomass (Fig 2). On one hand,
264 decreasing values from 0 mM to 7 mM Cr (III) corresponding to reductions in cfu
265 mL⁻¹ of 99.99%, cell viability of 37.1% and biomass of 92.94%, were also noted.
266 On the other hand, results of cell viability and biomass at 10 mM Cr(III) could
267 not be determined from FLU-CLSM images due to the formation of cellular
268 aggregates (Fig 2), but a minimal cfu mL⁻¹ value was achieved at this
269 concentration (Fig 1). Comparing these results with those obtained by different
270 phototrophic microorganisms, *Scenedesmus* sp. DE2009 and *Geitlerinema* sp.
271 DE2011 (Millach et al., 2015) and *Chroococcus* sp. PCC 9106 (Puyen et al.,

272 2017) and the yeast *Pichia stipitis* (Yilmazer and Saracoglu, 2008), it can be
273 demonstrated that *O. anthropi* DE2010 shows a great tolerance and resistance
274 to this metal at higher concentrations, even at 10 mM.

275

276 It is important to highlight that at 7 mM Cr(III) a high percentage of cells
277 remain alive (FLU-CLSM-IA method) despite minimal viable cells being
278 detected in plate counts. A possible explanation for this fact is the presence of
279 dormant cells (DC) in cultures contaminated with Cr(III) that are unable to grow
280 in plates. Lewis (2010) defined DC as a small sub-population of cells that
281 spontaneously enter a dormant, non-dividing, and resistant state. In this paper,
282 to study the role of DC, 24 h cultures of *O. anthropi* DE2010 grown at the
283 different tested Cr(III) concentrations were also spread onto LB agar without
284 metal (Fig 1A, cfu mL⁻¹ of total cells). These results are shown in Figure 2. The
285 number of DC increases as the metal concentration increases. Results were
286 significant at 5 mM Cr(III), with a relative ratio of dormant cells of 96.68%. This
287 can be seen even more clearly at 7 mM and 10 mM Cr(III), where the values
288 were 99.98% and 99.89%, respectively.

289

290 3.2 Changes in the EPS composition and production

291 The results concerning the biochemical analysis of EPS extracts from *O.*
292 *anthropi* DE2010 cultures grown with Cr(III) at different concentrations are
293 shown in Table 1. Changes in the EPS composition were observed as a
294 response to an increase in Cr(III) concentration. At lower concentrations of
295 Cr(III) (0-2 mM) the major EPS component is carbohydrates (88-89%), followed
296 by proteins (5-7%) and uronic acids (3-5%). However, at 5 mM Cr(III) the most

297 drastic metal effect was assessed. At this concentration, a considerable
298 increase of 37.88% in proteins was observed in contrast to a significant
299 reduction in 35.45% in carbohydrates, with as well as a slight decrease in uronic
300 acid content. Guibaud et al. (2003) demonstrated that the protein played a
301 major role in the complexation of metal ions and the number of binding sites
302 (alcohol, carboxyl and amino groups) for heavy metals. This change in the EPS
303 composition coincide with that described by Seng et al. (2005), who indicate
304 that the content of extracellular proteins increased when bacteria were grown in
305 adverse environmental conditions, and Yin et al. (2011), who report that the
306 presence of heavy metals resulted in a higher production of proteins than total
307 carbohydrates and uronic acids in the EPS. A decreasing value in
308 carbohydrates and proteins content is observed from 7 mM Cr(III), reaching a
309 similar value between both EPS components, although a slight increase was
310 observed at 10 mM, as result of cell aggregation. Regarding the uronic acids
311 content, in this study its role in *O. anthropi* DE2010 is less clear, although
312 different authors have been demonstrated the active role of this EPS
313 components in heavy metals chelation (Guibaud et al., 2005; Morillo Pérez et
314 al., 2008; Pal and Paul, 2008). The results showed in Table 1 demonstrated that
315 the concentration of uronic acid decreases from 0 to 5 mM and it is maintained
316 at higher Cr(III) concentrations, especially at 7 mM, when the carbohydrates
317 and proteins reach the lowest values. This data coincides with the formation of
318 cellular aggregates, which probably indicates that under these conditions,
319 uronic acid could contribute to chromium immobilization. Statistically significant
320 differences were found among all the metal concentrations assayed for
321 carbohydrates ($F= 271.15$) ($p <0.05$), proteins ($F= 471.14$) ($p <0.05$), and uronic

322 acids ($F= 130.28$) ($p < 0.05$). Using the Tukey multiple comparisons post-hoc
323 test, statistically significant differences ($p < 0.05$) were labelled in Table 1.

324

325 As regards the EPS production, a maximum increase in EPS
326 concentration in relation to cell concentration (7.82%) was obtained from 0 mM
327 (10.17%) to 5 mM (17.99%) Cr(III) coinciding with the increment in the protein
328 content. It is well known that the EPS generally contains high molecular weight
329 compounds with charged functional groups with adsorptive and adhesive
330 properties. These functional groups provide binding sites for heavy metals, as
331 demonstrated by Decho (1990) and more recently by other authors (More et al.,
332 2014; Yue et al., 2015). These results could explain the increase in total amount
333 of EPS at high concentrations of Cr(III) in *O. anthropi* DE2010.

334

335 As previously mentioned the most important cytotoxic effect of Cr(III) on
336 *O. anthropi* DE2010 is produced at 5 mM Cr(III). At this concentration, the
337 percentage of dead cells increases, provoking a drop in viability and biomass.
338 This is linked to an enhancement in the cellular synthesis of EPS by *O. anthropi*
339 DE2010 to protect the cells themselves from chromium toxicity, increasing, in
340 turn, the protein content and reducing the amount of carbohydrate and uronic
341 acids.

342

343 3.3 Removal capacity for Cr(III) of *O. anthropi* DE2010

344 Quantitative chemical data for the ability of *O. anthropi* DE2010 cells to
345 capture Cr(III) and the corresponding cellular uptake efficiency (q) are shown in
346 Table 2. These results indicate that there is a strong correlation between the

347 initial concentration of metal and the total of chromium taken from the cells,
348 reaching maximum values of removal of 40.83% at 7 mM and of 38.38% at 10
349 mM Cr(III). In addition, practically all the chromium removed was detected
350 inside the cells (cytoplasm), with values ranging between 97% and 99%,
351 depending on the metal concentration, with the rest of the metal being found in
352 the EPS extracts (1-3%). These results indicate that *O. anthropi* DE2010 is able
353 to immobilize chromium in both the EPS and the cytoplasm of the cells. The q
354 values also increase by increasing the initial metal concentration, and a
355 maximum value of 950 mg g⁻¹ was reached at 10 mM Cr(III). However, the most
356 notable increase in q values (590 mg g⁻¹) was observed from 5 mM to 7 mM
357 Cr(III). The highest q value reported here could be related to the cellular
358 aggregation monitored by CLSM imaging at 10 mM. This cellular organization is
359 more efficient to capture Cr(III) and to protect the cells from the cytotoxic effect
360 of the metal than the individual cells observed at the other metal concentrations.
361 Statistically significant differences for q values ($F= 1,596.45$) ($p < 0.05$) were
362 found among all the metal concentrations assayed. Using the Tukey multiple
363 comparisons *post-hoc* test, statistically significant differences ($p < 0.05$) were
364 observed from 5 mM Cr(III) and were included in Table 2. The specific removal
365 efficiency (q) of *O. anthropi* DE2010 is higher with respect to those obtained by
366 other prokaryotic: 120 mg g⁻¹ in *Sphaerotilus natans* (Solisio et al., 2000), 185
367 mg g⁻¹ in *Spirulina* sp. (Chojnacka et al., 2005), and 14.28 mg g⁻¹ in
368 *Rhodococcus opacus* (Calfa et al., 2008); and eukaryotic microorganisms:
369 11.30 mg g⁻¹ in *Saccharomyces cerevisiae* (Ksheminska et al., 2005), and
370 41.18 mg g⁻¹ in *Chlorella miniata* (Han et al., 2006). For this reason, it can be
371 considered that *O. anthropi* DE2010 show a higher capacity to sequestrate

372 chromium into the cells than other microorganisms and is, therefore, a great
373 candidate to bioremediate natural/artificial environments polluted with
374 chromium.

375

376 3.4 Cr(III) localization at cellular structure level

377 Electron microscopy imaging, SEM and TEM, have become a crucial tool
378 to visualize the ultrastructure of microorganisms (Rachel et al., 2010; Golding et
379 al., 2016; Solé et al., 2019) and coupled to an EDX detector to analyze the
380 elemental composition and its distribution in the samples (Maldonado et al.,
381 2011, Burgos et al., 2013; Coreño-Alonso et al., 2014; Millach et al., 2015;
382 Povedano-Priego et al., 2017). Although, these techniques have been very
383 useful in this regard, we applied for the first time the HAADF-STEM EDX for
384 locating, at the nanoscale level, heavy metals immobilized on different cellular
385 structures (Fig 3) evidencing whether specific microorganisms have a capacity
386 to immobilize them inside and/or outside cells in the same semi-fine section.

387

388 HAADF-STEM images exhibited discernible changes (changes in cell
389 morphology, an increase in bright inclusions, and the appearance of cellular
390 pleomorphic forms) in cells (Fig 3 A1, B1, and C1) due to the Cr (III), mainly at
391 the highest concentration (10 mM Cr(III)). EDX microanalyses shows that
392 chromium was not detectable, either externally or internally, in control samples
393 (0 mM) and up to 2 mM Cr(III). However, results obtained from 5 mM to 10 mM
394 indicated that the chromium signal was detected in the cellular structure.
395 Figures 3 A2, B2, and C2 show the STEM EDX chromium and phosphorous
396 composition line profiles that were taken along a bright inclusion from a semi-

397 fine section micrograph of *O. anthropi* DE2010 grown at 0, 5, and 10 mM Cr(III).
398 When compared the image intensity profile, with an increased signal in the
399 bright inclusion, it is clear that the most part of chromium was located in the
400 bright inclusions, together with an important phosphorous signal, and
401 secondarily on the cellular surface at the EPS level.

402

403 It is known that microorganisms play an important role in heavy metal
404 immobilization processes (Cheng et al., 2010) either at extra-cellular
405 (biosorption) and/or intra-cellular (bioaccumulation) levels (Guibaud et al., 2005;
406 Guine et al., 2006; Kothe et al., 2010; Puyen et al., 2012b; Esteve et al., 2013;
407 Burgos et al., 2013; Gutiérrez-Corona et al., 2016). On the other hand,
408 Velásquez and Dussan, (2009) demonstrated the relation between external
409 capture and passive process (live or dead cells) and internal bioaccumulation
410 and active process (live cells).

411

412 The present STEM EDX results demonstrate that *O. anthropi* DE2010
413 accumulates Cr(III) externally in EPS, and internally mainly in cytoplasmic
414 polyphosphate inclusions. Therefore, the chromium removal via sorption and
415 accumulation mechanisms in this bacterium was demonstrated using STEM
416 EDX.

417 **Conclusions**

418 Our results indicate that *O. anthropi* DE2010 is able to (i) tolerate and
419 resist the presence of high Cr(III) concentrations, showing changes in
420 production and EPS composition and increasing the number of dormant cells,

421 (ii) capture chromium at higher removal efficiencies, and (iii) immobilize it mainly
422 inside the cell in polyphosphate inclusions (bioaccumulation process), as well
423 as externally in the EPS (biosorption process).

424 As a future perspective, *O. anthropi* DE2010 could be considered as a
425 potential agent for bioremediation in Cr(III) contaminated environments since it
426 shows a great ability to tolerate and efficiently remove Cr(III) from polluted
427 cultures.

428 **Acknowledgments**

429 We express our thanks for the assistance of the UAB staff of Servei de
430 Micròscopia (<http://sct.uab.cat/microscopia/>) in particular Dr Alejandro Sanchez-
431 Chardi, the UAB Servei de Anàlisi Química (<http://sct.uab.cat/saq/>) especially Dr
432 Ignacio Villarroya. We also appreciate the help of Cristina Sosa and the
433 valuable comments and suggestions of Diana Gutiérrez, Irene López-Gómez,
434 and Estefania Solsona. This research was supported by the following grants:
435 FONCICYT (Ref. 95887), Ministerio de Economía y Competitividad (Refs.
436 CTQ2014-54553-C3-2-R and CGL2008-01891) and UAB postgraduate
437 scholarship to EV. The ICN2 is supported by the Severo Ochoa (MINECO,
438 grant no. SEV-2017-0706) and CERCA programmes.

439 **Conflicts of interest**

440 The authors declare no conflict of interests.

441 **References:**

442 Adav, S.S., Lee, D.J., 2008. Extraction of extracellular polymeric substances
443 from aerobic granule with compact interior structure. *J. Hazard. Mater.* 154
444 (1-3), 1120-1126. <https://doi.org/10.1016/j.jhazmat.2007.11.058>.

445 Balk, E.M., Tatsioni, A., Lichtenstein, A.H., Lau, J., Pittas, A.G., 2007. Effect of
446 chromium supplementation on glucose metabolism and lipids: a
447 systematic review of randomized controlled trials. *Diabetes Care* 30 (8),
448 2154–2163. <https://doi.org/10.2337/dc06-0996>.

449 Bhattacharya, A., Dey, P., Gola, D., Mishra, A., Malik, A., Patel, N., 2015.
450 Assessment of Yamuna and associated drains used for irrigation in rural
451 and peri-urban settings of Delhi NCR. *Environ. Monit. Assess.* 187, 4146.
452 <https://doi.org/10.1007/s10661-014-4146-2>.

453

454 Bradford, M., 1976. A rapid and sensitive method for the quantification of
455 microgram quantities of protein utilizing the principle of protein– dye
456 binding. *Anal. Biochem.* 72, 248–254. [https://doi.org/10.1016/0003-
457 2697\(76\)90527-3](https://doi.org/10.1016/0003-2697(76)90527-3).

458

459 Burgos, A., Maldonado, J., De los Rios, A., Solé, A., Esteve, I., 2013. Effect of
460 copper and lead on two consortia of phototrophic microorganisms and
461 their capacity to sequester metals. *Aquat. Toxicol.* 140–141, 324–336.
462 <https://doi.org/10.1016/j.aquatox.2013.06.022>.

463

464 Burton, K., 1956. A study of the conditions and mechanism of the
465 diphenylamine reaction for the colorimetric estimation of deoxyribonucleic
466 acid. *Bioch.* 62, 315-323.

467

468 Calfa, BA., Torem, ML., 2008. On the fundamentals of Cr(III) removal from
469 liquid streams by a bacterial strain. Miner. Eng. 21 (1), 48-54.

470 <https://doi.org/10.1016/j.mineng.2007.08.001>.

471

472 Chakravarty, R., Banerjee, P.C., 2012. Mechanism of cadmium binding on the
473 cell wall of an acidophilic bacterium. Bioresour. Technol. 108, 176-183.

474 <https://doi.org/10.1016/j.biortech.2011.12.100>.

475

476 Cheng, Y., Yan, F., Huang, F., Chu, W., Pan, D., Chen, Z., Zheng, J., Yu, M.,
477 Lin Z., Wu, Z., 2010. Bioremediation of Cr (VI) and immobilization as Cr
478 (III) by *Ochrobactrum anthropi*. Environ. Sci. Technol. 44 (16), 6357–6363.

479 <https://doi.org/10.1021/es100198v>.

480

481 Chojnacka, K., Chojnacki, A., Górecka, H., 2005. Biosorption of Cr³⁺, Cd²⁺ and
482 Cu²⁺ ions by blue–green algae *Spirulina* sp.: kinetics, equilibrium and the
483 mechanism of the process. Chemosphere 59 (1), 75-84.

484 <https://doi.org/10.1016/j.chemosphere.2004.10.005>.

485

486 Coreño-Alonso, A., Solé, A., Diestra, E., Esteve, I., Gutiérrez-Corona, J.F.,
487 Reyna López, G.E., Fernández, F.J., Tomasini, A., 2014. Mechanisms of
488 interaction of chromium with *Aspergillus niger* var *tubingensis* strain Ed8.
489 Bioresour. Technol. 158, 188–192.

490 <https://doi.org/10.1016/j.biortech.2014.02.036>.

491

492 Decho, A.W., 1990. Microbial exopolymer secretions in ocean environments:
493 Their role(s) in food webs and marine processes. *Oceanogr. Mar. Biol.*
494 *Annu. Rev.* 28 (7), 73-153. <https://doi.org/10.1016/j.carbpol.2018.04.126>.
495

496 Dubois, M., Gilles, K.A., Hamilton, J.K., Rebers, P.A., Smith, F., 1956.
497 Colorimetric method for determination of sugars and related substances.
498 *Anal. Chem.* 28 (3), 350-356.
499

500 Esteve, I., Maldonado, J., Burgos, A., Diestra, E., Burnat, M., Solé, A., 2013.
501 Confocal laser scanning and electron microscopic techniques as powerful
502 tools for determining the in vivo effect and sequestration capacity of lead
503 in cyanobacteria. *Cyanobacteria: Ecology, Toxicology and Management*.
504

505 Evert, A.B., Boucher, J.L.B., Cypress, M., Dunbar, S.A., Franz, M.J., Mayer-
506 Davis, E.J., Neumiller, J.J., Nwankwo, R., Verdi, C.L., Urbanski, P., Yancy,
507 W.S., 2013. Nutrition therapy recommendations for the management of
508 adults with diabetes. *Diabetes Care* 36 (11), 3821-3842.
509 <http://doi.org/10.2337/dc13-2042>.
510

511 Feng, S., Mai, B., Wei, G., Wang, X., 2012. Genotoxicity of the sediments
512 collected from Pearl River in China and their polycyclic aromatic
513 hydrocarbons (PAHs) and heavy metals. *Environ. Monit. Assess.* 184,
514 5651–5661. <http://doi.org/10.1007/s10661-011-2369-z>.
515

516 Figgitt, M., Newson, R., Leslie, I.J., Fisher, J., Ingham, E., Case, C.P., 2010.
517 The genotoxicity of physiological concentrations of chromium (Cr(III) and
518 Cr(VI)) and cobalt (Co(II)): An in vitro study. *Mutat. Res.* 688, 53–61.
519 <http://doi.org/10.1016/j.mrfmmm.2010.03.008>.
520

521 Francisco, R., de Abreu, P., Plantz, B.A., Schlegel, V.L., Carvalho, R.A.,
522 Vasconcelos-Morais, P., 2011. Metal-induced phosphate extracellular
523 nanoparticulate formation in *Ochrobactrum tritici* 5bv11. *J. Hazard. Mater.*
524 198, 31-39. <https://doi.org/10.1016/j.jhazmat.2011.10.005>.
525

526 Golding, C.G., Lamboo, L.L., Beniac, D.R., Booth, T.F., 2016. The scanning
527 electron microscope in microbiology and diagnosis of infectious disease.
528 *Sci. Rep.* 6, 26516. <https://doi.org/10.1038/srep26516>
529

530 Greenwood, N.N., Earnshaw, A., 1997. *Chemistry of the elements*, 2nd ed.
531 Butterworth–Heinemann, Oxford.
532

533 Guibaud, G., Tixier, N., Bouju, A., Baudu, M., 2003. Relation between
534 extracellular polymers' composition and its ability to complex Cd, Cu and
535 Pb. *Chemosphere* 52(10),1701-1710. [https://doi.org/10.1016/S0045-](https://doi.org/10.1016/S0045-6535(03)00355-2)
536 [6535\(03\)00355-2](https://doi.org/10.1016/S0045-6535(03)00355-2).
537

538 Guibaud, G., Comte, S., Bordas, F., Dupuy, S., Baudu, M., 2005. Comparison
539 of the complexation potential of extracellular polymeric substances (EPS),
540 extracted from activated sludges and produced by pure bacteria strains,

541 for cadmium, lead and nickel. Chemosphere 59 (5), 629-638.
542 <http://doi.org/10.1016/j.chemosphere.2004.10.028>.
543
544 Guiné, V., Spadini, L., Sarret, G., Muris, M., Delolme, C., Gaudet, J.P., Martins,
545 J.M.F., 2006. Zinc sorption to three gram-negative bacteria: combined
546 titration, modeling, and EXAFS study. Environ. Sci. Technol. 40 (6),1806–
547 1813. <https://doi.org/10.1021/es050981l>.
548
549 Gupta, S.K., Chabukdhara, M., Kumar, P., Singh, J., Bux, F., 2014. Evaluation
550 of ecological risk of metal contamination in river Gomti, India: a
551 biomonitoring approach. Ecotoxicol. Environ. Saf. 110, 49–55.
552 <http://doi.org/10.1016/j.ecoenv.2014.08.008>.
553
554 Gutiérrez-Corona, J.F., Romo-Rodríguez, P., Santos-Escobar, F., Espino-
555 Saldaña, A.E., Hernández-Escoto, H., 2016. Microbial interactions with
556 chromium: basic biological processes and applications in environmental
557 biotechnology. World. J. Microbiol. Biotechnol. 32 (12), 191.
558 <https://doi.org/10.1007/s11274-016-2150-0>.
559
560 Han, X., Wong, YS., Tam, NFY., 2006. Surface complexation mechanism and
561 modeling in Cr(III) biosorption by a microalgal isolate, *Chlorella miniate*. J.
562 Colloid Interface Sci. 303 (2), 365-371.
563 <https://doi.org/10.1016/j.jcis.2006.08.028>.
564

565 Ihsanullah, Abbas, A., Al-Amer, A.M., Laoui, T., Al-Marri, M.J., Nasser, M.S.,
566 Khraisheh, M., Atieh, M.A., 2016. Heavy metal removal from aqueous
567 solution by advanced carbon nanotubes: Critical review of adsorption
568 applications. Sep. Purif. Technol. 157 (8), 141-161.
569 <https://doi.org/10.1016/j.seppur.2015.11.039>.

570 Kimbrough, D.E., Cohen, Y., Winer, A.M., Creelman, L., Mabuni, C., 1999. A
571 Critical assessment of chromium in the environment. Crit. Rev. Environ.
572 Sci. Technol. 29 (1), 1-46. <https://doi.org/10.1080/10643389991259164>.

573 Kintner, P.K., van Buren, J.P., 1982. Carbohydrate Interference and Its
574 Correction in pectin analysis using the *m*-Hydroxydiphenyl method. J.
575 Food Sci. 47, 756-759. [https://doi.org/10.1111/j.1365-
576 2621.1982.tb12708.x](https://doi.org/10.1111/j.1365-2621.1982.tb12708.x).

577 Kothe, E., Dimkpa, C., Haferburg, G., Schmidt, A., Schmidt, A., Schütze, E.,
578 2010. Streptomycete heavy metal resistance: extracellular and intracellular
579 mechanisms, in: Sherameti, I., Varma, A., (Eds.). Soil Heavy Metals, Soil
580 Biology. https://doi.org/10.1007/978-3-642-02436-8_10.

581

582 Ksheminska, H., Fedorovych, D., Babyak, L., Yanovych, D., Kaszycki, P.,
583 Koloczek, H., 2005. Chromium (III) and (VI) tolerance and bioaccumulation
584 in yeast: A survey of cellular chromium content in selected strains of
585 representative genera. Process Biochemistry. 40, 1565-1572.
586 <https://doi.org/10.1016/j.procbio.2004.05.012>.

587 Kusiak, R.A., Ritchie, A.C., Springer, J., Muller, J., 1993. Mortality from stomach
588 cancer in Ontario miners. Br. J. Med. 50,117–126.

589

590 Lewis K., 2010. Persister Cells. *Annu. Rev. Microbiol.* 64, 357–372.

591

592 Liu, H., Fang, H.P.H., 2002. Extraction of extracellular polymeric substances
593 (EPS) of sludges. *J. Biotechnol.* 95 (3), 249-256.

594 [https://doi.org/10.1016/S0168-1656\(02\)00025-1](https://doi.org/10.1016/S0168-1656(02)00025-1).

595

596 Maldonado, J., Diestra, E., Domènech, A.M., Villagrasa, E., Puyen, Z.M,
597 Esteve, I., Solé, A., 2010. Isolation and identification of a bacterium with
598 high tolerance to lead and copper from a marine microbial mat in Spain.
599 *Ann. Microbiol.* 60 (1), 113-120. [https://doi.org/10.1007/s13213-010-0019-](https://doi.org/10.1007/s13213-010-0019-2)

600 [2](https://doi.org/10.1007/s13213-010-0019-2).

601

602 Maldonado, J., Solé, A., Puyen, Z.M., Esteve, I., 2011. Selection of
603 bioindicators to detect lead pollution in Ebro delta microbial mats, using
604 high-resolution microscopic techniques. *Aquat.Toxicol.* 104, 135–144.

605 <http://doi.org/10.1016/j.aquatox.2011.04.009>.

606

607 Manahan, S.E., 2009. *Environmental chemistry*, 9th edition. Boca Raton: CRC
608 Press, Florida.

609

610 Millach, L., Solé, A., Esteve, I., 2015. Role of *Geitlerinema* sp. DE2011 and
611 *Scenedesmus* sp. DE2009 as bioindicators and immobilizers of chromium
612 in a contaminated natural environment. *BioMed Research International*,
613 Article ID 519769. <https://doi.org/10.1155/2015/519769>.

614

615 Millonig, G., 1961. A modified procedure for lead staining of thin sections. J
616 Biophys. Biochem. Cytol. 11 (3), 736–739.

617

618 Mohana, D., Pittman, C.U., 2006. Activated carbons and low cost adsorbents
619 for remediation of tri- and hexavalent chromium from water. J. Hazard.
620 Mater. 137 (2), 762-811. <https://doi.org/10.1016/j.jhazmat.2006.06.060>.

621

622 More, T.T., Yadav, J.S.S., Yan, S., Tyagi, R.D., Surampalli. R.Y., 2014.
623 Extracellular polymeric substances of bacteria and their potential
624 environmental Applications. J. Environ. Manage. 144, 1-25.
625 <https://doi.org/10.1016/j.jenvman.2014.05.010>.

626

627 Morillo Pérez, J.A., García-Ribera, R., Quesada, T., Aguilera, M., Ramos-
628 Cormenzana, A., Monteoliva-Sánchez M., 2008. Biosorption of heavy
629 metals by the exopolysaccharide produced by *Paenibacillus jamilae*.
630 World J. Microbiol. Biotechnol. 24, 2699. [https://doi.org/10.1007/s11274-](https://doi.org/10.1007/s11274-008-9800-9)
631 [008-9800-9](https://doi.org/10.1007/s11274-008-9800-9).

632

633 Nziguheba, G., Smolders, E., 2008. Inputs of trace elements in agricultural soils
634 via phosphate fertilizers in European countries. Sci. Total. Environ. 390
635 (1), 53-57. <https://doi.org/10.1016/j.scitotenv.2007.09.031>.

636

637 Pal, A. and Paul, A.K., 2008. Microbial extracellular polymeric substances:
638 central elements in heavy metal bioremediation. *Indian J. Microbiol.* 48,
639 49-64.
640

641 Povedano-Priego, C., Martín-Sánchez, I., Jroundi, F., Sánchez-Castro, I.,
642 Merroun, M.L., 2017. Fungal biomineralization of lead phosphates on the
643 surface of lead metal. *Miner. Eng.* 106, 46–54.
644 <https://doi.org/10.1016/j.mineng.2016.11.007>.
645

646 Puyen, Z.M., Villagrasa, E., Maldonado, J., Esteve, I., Solé, A., 2012a. Viability
647 and biomass of *Micrococcus luteus* DE2008 at different salinity
648 concentrations determined by specific fluorochromes and CLSM-image
649 analysis. *Curr. Microbiol.* 64 (1), 75-80. [https://doi.org/10.1007/s00284-](https://doi.org/10.1007/s00284-011-0033-z)
650 [011-0033-z](https://doi.org/10.1007/s00284-011-0033-z).
651

652 Puyen, Z.M., Villagrasa, E., Maldonado, J., Diestra, E., Esteve, I., Solé, A.,
653 2012b. Biosorption of lead and copper by heavy-metal tolerant
654 *Micrococcus luteus* DE2008. *Bioresour. Technol.* 126, 233–237.
655 <https://doi.org/10.1016/j.biortech.2012.09.036>.
656

657 Puyen, Z. M., Villagrasa, E., Millach, L., Esteve, I., Maldonado, J., Solé, A.,
658 2017. Multi-approach microscopy techniques to evaluate the cytotoxic
659 effect of chromium(III) on the cyanobacterium *Chroococcus* sp. PCC 9106.
660 in: Menendez-Vilas, A. (Ed). *Microscopy and imaging science: practical*

661 approaches to applied research and education. Formatex Research
662 Center, Badajoz, pp. 602-609.
663
664 Rachel, R., Meyer, C., Klingl, A., Gürster, S., Heimerl, T., Wasserburger, N.,
665 Burghardt, T., Küper, U., Bellack, A., Schopf, S., Wirth, R., Huber, H.,
666 Wanner G., 2010. Chapter 3 - Analysis of the Ultrastructure of Archaea by
667 Electron Microscopy, Editor(s): Thomas Müller-Reichert. Methods in Cell
668 Biology. Academic Press. Vol(96), pages: 47-69.
669 [https://doi.org/10.1016/S0091-679X\(10\)96003-2](https://doi.org/10.1016/S0091-679X(10)96003-2).
670
671 Sachidanandham, R., Yew-Hoong Gin, K., 2009. A dormancy state in nonspore-
672 forming bacteria. Appl. Microbiol. Biotechnol. 81, 927.
673 <https://doi.org/10.1007/s00253-008-1712-y>.
674
675 Schneider, C.A., Rasband, W.S., Eliceiri, K.W., 2012. NIH Image to ImageJ: 25
676 years of Image Analysis. Nat. Methods 9 (7), 671–675.
677
678 Shanker, A.K., Cervantes, C., Loza-Tavera, H., Avudainayagam, S., 2005.
679 Chromium toxicity in plants. Environ. Int. 31 (5), 739-753.
680 <https://doi.org/10.1016/j.envint.2005.02.003>.
681
682 Sheng, G.P., Yu, H.Q., Yue, Z.B., 2005. Production of extracellular polymeric
683 substances from *Rhodopseudomonas acidophila* in the presence of toxic
684 substances. Appl. Microbiol. Biotechnol. 69 (2), 216-222.
685 <https://doi.org/10.1007/s00253-005-1990-6>.

686

687 Sole, A., Calvo, M.A., Lora, M.J., Sánchez-Chardi, A., 2019. Chapter 12 -
688 Electron microscopy techniques applied to bioremediation and
689 biodeterioration studies with molds. Editor(s): Araceli Tomasini
690 Campocosio and Hector Hugo Leon Santiesteban. Fungal bioremediation:
691 fundamentals and applications. CRC press. pages: 354.
692 <https://doi.org/10.1201/9781315205984>.

693

694 Solisio, C., Lodi, A., Converti, A., Del Borghi, M., 2000. The effect of acid pre-
695 treatment on the biosorption of chromium(III) by *Sphaerotilus natans* from
696 industrial wastewater. Water Res. 34 (12), 3171-3178.
697 [https://doi.org/10.1016/S0043-1354\(00\)00059-2](https://doi.org/10.1016/S0043-1354(00)00059-2).

698

699 Srivastava, S., Thakur, I.S., 2007. Evaluation of biosorption potency of
700 *Acinetobacter* sp. for removal of hexavalent chromium from tannery
701 effluent. Biodegradation 18 (5), 637–646. [https://doi.org/10.1007/s10532-](https://doi.org/10.1007/s10532-006-9096-0)
702 [006-9096-0](https://doi.org/10.1007/s10532-006-9096-0).

703

704 Suwalsky, M., Castro, R., Villena, F., Sotomayor, C.P., 2008. Cr(III) exerts
705 stronger structural effects than Cr(VI) on the human erythrocyte
706 membrane and molecular models. J. Inorg. Biochem. 102, 842–849.
707 <https://doi.org/10.1016/j.jinorgbio.2007.11.020>.

708

709 Velásquez, L., Dussan, J., 2009. Biosorption and bioaccumulation of heavy
710 metals on dead and living biomass of *Bacillus sphaericus*. J. Hazard.
711 Mater. 167 (1-3), 713–6. <https://doi.org/10.1016/j.jhazmat.2009.01.044>.
712

713 Villagrasa, E., Ferrer-Miralles, N., Millach, L., Obiol, A., Creus, J., Esteve, I.,
714 Solé, A., 2019. Morphological responses to nitrogen stress deficiency of a
715 new heterotrophic isolated strain of Ebro Delta microbial mats.
716 Protoplasma. 256 (1), 105-116. [https://doi.org/10.1007/s00709-018-1263-](https://doi.org/10.1007/s00709-018-1263-8)
717 [8](https://doi.org/10.1007/s00709-018-1263-8).
718

719 Xu, X., Zhao, Y., Zhao, X., Wang, Y., Deng, W., 2014. Sources of heavy metal
720 pollution in agricultural soils of a rapidly industrializing area in the Yangtze
721 Delta of China. Ecotoxicol. Environ. Saf. 108, 161–167.
722 <https://doi.org/10.1016/j.ecoenv.2014.07.001>.
723

724 Yilmazer, P., Saracoglu, N., 2009. Bioaccumulation and biosorption of
725 copper(II) and chromium(III) from aqueous solutions by *Pichia stipitis*
726 yeast. J. Chem. Technol. Biotechnol. 84, 604–610.
727 <https://doi.org/10.1002/jctb.2088>.
728

729 Yin, Y., Hu, Y., Xiong, F., 2011. Sorption of Cu (II) and Cd (II) by extracellular
730 polymeric substances (EPS) from *Aspergillus fumigates*. Int. Biodeterior.
731 Biodegrad., 65, 1012-1018. <https://doi.org/10.1016/j.ibiod.2011.08.001>.
732

733 Yue, Z.B., Li, Q., Li, C.C., Chen, T.H., Wang, J., 2015. Component analysis and
734 heavy metal adsorption ability of extracellular polymeric substances (EPS)
735 from sulfate reducing bacteria. *Bioresour. Technol.* 194, 399-402.
736 <https://doi.org/10.1016/j.biortech.2015.07.042>.

Table 1 Variations in the production and composition of the EPS of *Ochrobactrum anthropi* DE2010

Cr(III) concentration (mM)	[DNA] (ppm)	[Carbohydrates] (ppm)	[Proteins] (ppm)	[Uronic acids] (ppm)	[Cell co] (ppm)
0	0.28 ± 0.02	148.12 ± 4.94	9.98 ± 0.71	8.40 ± 0.71	1640 ± 8
0.5	0.58 ± 0.02**	150.50 ± 2.09	11.95 ± 0.52	6.30 ± 0.34**	1580 ± 1
2	0.91 ± 0.02****	133.28 ± 5.44	11.89 ± 0.88	6.02 ± 0.20**	1505 ± 1
5	1.20 ± 0.02****+\$	45.36 ± 0.95****+\$	39.37 ± 0.44****	1.35 ± 0.16****+\$	485 ± 24
7	0.87 ± 0.03 ^a ****&&	12.46 ± 0.15****\$\$&&	10.72 ± 1.16&&	0.97 ± 0.07****\$\$&&	255 ± 4.6
10	1.27 ± 0.03 ^a ****\$\$ [^]	17.36 ± 0.24****\$\$&&	12.47 ± 4.36* [^] &&	1.37 ± 0.26****\$\$ [^]	220 ± 4.7

^a values out of range indicating cell lysis due to Cr(III)

Data are expressed as mean ±SD (n= 15)

p* <0.05; *p* <0.01: 0.5, 2, 5, 7 and 10 mM treatments vs. control (0 mM)

+*p* <0.05; ++*p* <0.01: 2, 5, 7, and 10 mM treatments vs. 0.5 mM treatment

\$\$*p* <0.01: 5, 7, and 10 mM treatments vs. 2 mM treatment

&&*p* <0.01: 7, and 10 mM treatments vs. 5 mM treatment

[^]*p* <0.01: 7 treatment mM vs. 10 mM treatment

Table 2 Removal Cr(III) capacity by *Ochrobactrum anthropi* DE2010 in the EPS and cytoplasm

Cr(III) concentration (mM)	Culture Medium (C _i) (mg L ⁻¹)	Supernatant (C _f) (mg L ⁻¹)	Total metal removed (mg L ⁻¹)	Metal in EPS (mg L ⁻¹)	Metal in cells without EPS (mg L ⁻¹)	Dry weight (m) (g)
0	0	0	0	0	0	0.033 ± 0.001
0.5	24 ± 1.15	16.50 ± 0.57	7.50 ± 1.73	0.20 ± 0.006	7.30 ± 1.71	0.031 ± 0.001
2	99 ± 2.30	83.50 ± 1.73	15.50 ± 4.04	0.21 ± 0.010	15.29 ± 4.03	0.030 ± 0.001
5	257 ± 2.31	199 ± 3.46	58 ± 1.15	0.52 ± 0.015	57.48 ± 1.14	0.010 ± 0.001
7	360 ± 3.46	212.50 ± 2.89	147.50 ± 0.58	1.85 ± 0.045	145.15 ± 0.54	0.005 ± 0.001
10	495 ± 9.24	305 ± 8.08	190 ± 8.99	5.10 ± 0.095	184.90 ± 0.90	0.004 ± 0.001

Data are expressed as mean ±SD (n= 4)

* $p < 0.05$, ** $p < 0.01$: 0.5, 2, 5, 7 and 10 mM treatments vs. control (0 mM)

++ $p < 0.01$: 2, 5, 7, and 10 mM treatments vs. 0.5 mM treatment

\$\$ $p < 0.01$: 5, 7, and 10 mM treatments vs. 2 mM treatment

&& $p < 0.01$: 7, and 10 mM treatments vs. 5 mM treatment

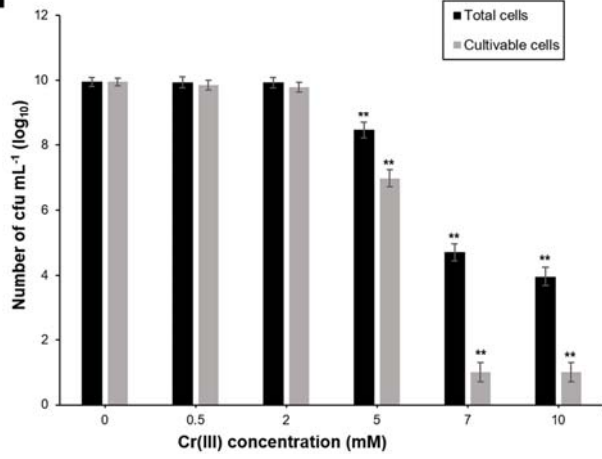
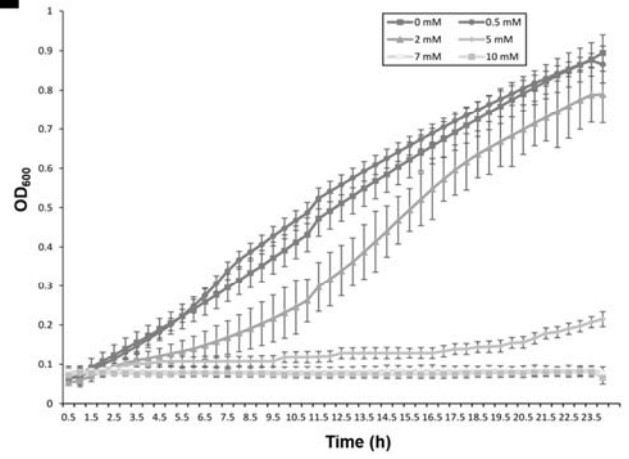
^^ $p < 0.01$: 7 treatment mM vs. 10 mM treatment

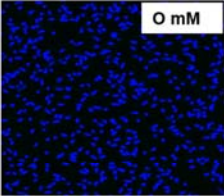
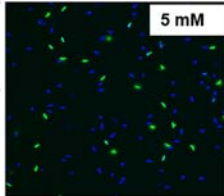
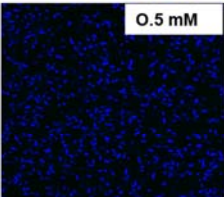
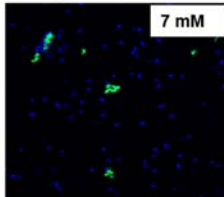
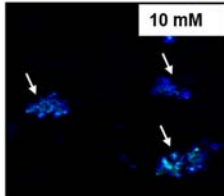
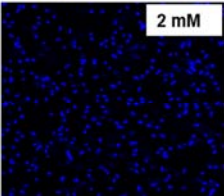
1 **Figure captions**

2 **Figure 1** Growth parameters of *O. anthropi* DE2010 growing at different Cr(III)
3 concentrations. **(A)** Plate counts (cfu mL⁻¹) from total (grown on LB agar without
4 Cr(III)) and cultivable cells (grown on LB agar with Cr(III)). $p < 0.01^{**}$: 0.5, 2, 5,
5 7, and 10 mM treatments vs. control (0 mM). **(B)** Growth curves at different
6 Cr(III) working concentrations. In both graphics, data are expressed as mean
7 \pm SD (n= 8).

8 **Figure 2** Cytotoxic effect of Cr(III) in the *O. anthropi* DE2010 cultures. $*p < 0.05$,
9 $**p < 0.01$: 0.5, 2, 5, and 7 mM treatments vs. control (0 mM). $+p < 0.05$, $++p <$
10 0.01 : 2, 5, and 7 mM treatments vs. 0.5 mM treatment. $$$p < 0.01$: 5, and 7 mM
11 treatments vs. 2 mM treatment. Data are expressed as mean \pm SD (n= 20).

12 **Figure 3** HAADF STEM analysis of *O. anthropi* DE2010 cells grown at 0 **(A)**, 5
13 **(B)**, and 10 mM **(C)** of Cr(III). **(A1, B1, and C1)** HAADF STEM imaging, and
14 **(A2, B2, and C2)** HAADF STEM-EDX analysis and line profiles (250 nm) along
15 the black line showing the intensity of the HAADF image (top), phosphorus
16 (middle) and chromium (bottom).

A**B**

 0 mM	Cr (III) Concentration (mM)	Live cells (%)	Dead cells (%)	Biomass (mg C/ cm ³)	DC (%)	 5 mM
	0	99.82 ± 9.13	0.18 ± 0.006	46.76 ± 8.57	0	
 0.5 mM	0.5	99.75 ± 8.17	0.25 ± 0.009	32.69 ± 5.27*	17.04 ± 1.06*	 7 mM
	2	99.55 ± 8.11*	0.45 ± 0.012*	27.40 ± 3.83**	27.95 ± 1.78****	
	5	74.21 ± 3.09****SS	25.79 ± 1.03****SS	6.38 ± 1.56****SS	96.68 ± 2.43****SS	 10 mM
 2 mM	7	62.72 ± 0.85****SS	37.28 ± 0.43****SS	3.30 ± 0.40****SS	99.98 ± 2.76****SS	
	10	ND	ND	ND	99.89 ± 3.34****SS	

