



This is the **accepted version** of the journal article:

Ehlers, Bodil K.; Berg, Matty P.; Staudt, Michael; [et al.]. «Plant secondary compounds in soil and their role in belowground species interactions». Trends in Ecology and Evolution, Vol. 35, Issue 8 (August 2020), p. 716-730. DOI 10.1016/j.tree.2020.04.001

This version is available at https://ddd.uab.cat/record/257070

under the terms of the GO BY-NC-ND license

4

- 1 Plant secondary compounds in soil and their role in belowground species interactions
- 2 Ehlers BK¹, Berg MP^{2,3}, Staudt M⁴, Holmstrup M¹, Glasius M⁵, Ellers J², Tomiolo S¹, Bjerregaard
- 3 R^5 , Slotsbo S^1 , and Penuelas $J^{6,7}$.
- 5 Full contact details for all authors including institution, city, country, and postcode.
- 6 1. Department of Bioscience, Aarhus University, Vejlsøvej 25, 8600 Silkeborg, Denmark
- 7 2. Community and Conservation Ecology Group, Groningen Institute of Evolutionary Life
- 8 Sciences, University of Groningen, Nijenborgh 7, 9747 AG Groningen, The Netherlands.
- 9 3. Department of Ecological Science, Vrije Universiteit Amsterdam, De Boelelaan 1085, 1081 HV
- 10 Amsterdam, The Netherlands
- 4. CEFE, CNRS, Univ Montpellier, Univ Paul Valéry Montpellier 3, EPHE, IRD,
- 12 1919 Route de Mende, 34293 Montpellier, France
- 5. Department of Chemistry and Interdisciplinary Nanoscience Center, Langelandsgade 140, 8000
- 14 Århus, Denmark.
- 15 6. CSIC, Global Ecology Unit CREAF-CSIC-UAB, 08193 Bellaterra, Catalonia, Spain.
- 7. CREAF, 08193 Cerdanyola del Vallès, Catalonia, Spain
- 18 Twitter handles for authors @SaraTomiolo
- 20 Keywords: Chemical ecology, Soil functioning, Climate change, Soil organisms, Aboveground-
- 21 belowground interactions
- 23 Abstract

17

19

22

- 25 Knowledge about the effect of plant secondary compounds (PSCs) on belowground interactions in
- 26 the more diffuse community of species living outside the rhizosphere is sparse compared to what
- 27 we know about how PSCs affect aboveground interactions. We illustrate here that PSCs from foliar
- 28 tissue, root exudates, and leaf litter effectively influence such belowground plant-plant, plant-
- 29 microorganism, and plant-soil invertebrate interactions. Climatic factors can induce PSC production
- and select for different plant chemical types. Climate change can therefore alter both quantitative
- and qualitative PSC production, and how these compounds move in the soil. This can change the

- 32 soil chemical environments, with cascading effects on both the ecology and evolution of
- 33 belowground species interactions, and ultimately soil functioning.

Plant chemicals mediate species interactions

Secondary compounds in plants are major contributors to the chemical diversity of nature. They are formed from **primary metabolites** (see Glossary) in specific pathways and omnipresent in plants. The distribution of **plant secondary compounds** (PSCs) is heterogeneous across the plant kingdom, and these compounds exhibit extensive variation both among and within species [1]. Though many PSCs also play pivotal roles in the plant's primary metabolic processes (such as growth or development), the large majority are known to serve as infochemicals mediating interactions between plants and their biotic and abiotic environments. PSCs can profoundly modify the abiotic environment as they are released into the atmosphere, soil, and water of an ecosystem, which in turn determines the community structure of the biotic constituents. For example, PSCs that accumulate in stems and foliage of the vegetation can promote bushfires, that profoundly alter the physicochemical properties of soil and groundwater [2]. Many PSCs are also involved in ecologically crucial interactions for plants, as they can form chemical barriers against pathogens and herbivores, attract pollinators and predators of herbivores, and help disperse fruit and seeds [3]. To fine-tune the release of PSCs, plants "eavesdrop" on their neighbours and increase their own defences when they sense volatiles released from neighbouring plants damaged by leaf herbivores [4].

PSC-mediated interactions between plants and their associated organisms are best-documented for aboveground compounds associated with airborne transport and signalling [3, 5, 6]. PSCs, however, are also omnipresent in the soil, where they mediate interactions with soil organisms and neighbouring plants. Compared to aboveground interactions, these soil PSCs represent a "hidden" link between organisms. Recent reviews of PSCs and belowground interactions [7-9] have focused on PSCs in root exudates and on species interactions, mainly in the **rhizosphere**. However, less attention has been given to the impact of PSCs on the broader and more diffuse community of associated species further away from roots, which also includes the effects of PSC from foliar tissue and litter decomposition. Climate change may alter both the quantitative and qualitative production of PSCs, which could cascade to affect belowground PSC-mediated community interactions. Given the vital role PSCs play aboveground, we expect PSCs to be equally important to belowground plant-organism interactions.

Here, we focus on four major groups of PSCs: terpenes, flavonoids, glucosinolates, and alkaloids (Table 1) that enter the soil matrix. We discuss their origin and fate in the soil and their ecological

and evolutionary impacts on belowground interactions. Release of PSCs by plants alters soil chemistry, and neighbouring species can either tolerate, adapt to or escape the chemical environment this imposes on them. However, the production, release, and transport of PSCs by plants are greatly affected by local conditions, such as microclimate, microbes, herbivore and pathogen pressure. Therefore, changes in climate will affect the quantitative and qualitative production and release of PSCs both directly through changes in temperature and humidity but also indirectly through potential concomitant changes in herbivore and pathogen pressure. As the effects of PSCs on neighbour species are often both compound-specific and dose-specific, there is the potential for cascading effects of PSC-driven alteration of soil chemistry on belowground interactions. So far, however, very little is known regarding such cascading effects in plant-soil community interactions. To encourage more studies on this topic, we outline knowledge gaps and describe best practices for sampling and detecting PSCs in soil, to designate directions for future studies advancing this field of research.

80 Origin of PSCs in soils

How do PSCs enter the soil? Those formed in roots can be excreted directly into the soil from living and deteriorating tissues [10]. PSCs formed in aboveground organs are deposited into the soil from belowground organs following internal transport [11] or during wet and dry deposition from the atmosphere after volatilization and further chemical reactions [12-14]. PSCs can also be washed from vegetation surfaces with water intercepted during precipitation [15], or they may outgas and leach from fresh and decomposing litter on the soil surface and in the upper horizons [16, 17]. Many plant species accumulate large quantities of PSCs in secretory organs that are either external, such as trichomes, or internal, such as oil blisters or resin ducts, and a strong discontinuous release is expected during the decline and decomposition of these organs.

Concentrations of PSCs can be high and persistent in the soil. For instance, mono- and sesquiterpenes released from pine needles, litter and roots are detectable in forest soil throughout the year [18] and can attain high concentrations especially in topsoil but also lower mineral soil, being highest in late summer, and in the winter during snow cover [19]. Glucosinolates, produced by all cruciferous plants, are hydrolysed in the soil to highly bioactive compounds, notably isothiocyanate, nitriles, and thiocyanate. The degradation and hydrolyzation of glucosinolates depend on the presence of the extracellular enzyme myrosinase [20]. Glucosinolates are water-soluble and can more easily be washed out from soil, but isothiocyanates are hydrophobic and adsorb strongly to organic matter [20, 21], and may thus persist in the soil for periods of time. The isothiocyanate sinigrin, for example, is

persistent in the soil of forests invaded by *Alliaria petiolata* (garlic mustard), where even low concentrations suppress germination and growth of arbuscular mycorrhiza. This suppression disrupt the mycorrhizal association with native plant, negatively affecting the native plant's resource uptake [22]. The persistence and slow degradation of many PSCs in soil emphasize the ecological and evolutionary potential pressure PSCs can impose on belowground species.

PSC production under climate change

The production of PSCs improves the ability of plants to cope with different abiotic [23] and biotic stressors [5, 6]. Various abiotic factors, such as CO₂ and temperature, affect PSC release by up- or downregulating their production, although the specific PSC compounds that are affected may differ among factors. For example, in northern hemisphere trees, high CO₂ levels generally increase the foliar concentration of phenolic compounds such as flavonoids and tannins, and decrease the production of terpenoid compounds, whereas elevated temperatures lead to the opposite trend [24-26]. Drought can both increase and decrease PSC production dependent on the severity of drought and on the compound. Terpenoids and phenolics in temperate and Mediterranean trees and herbs increased under moderate drought stress but decreased under severe drought [25, 27, 28]. Although severe drought overall decreased total terpenoid emission, some specific mono- and sesquiterpenes either increased or remained unaffected, suggesting an adaptive role of specific compounds in mitigating abiotic stress [27, 28].

In addition to environmental effects on the production of specific compounds or compound classes, genetically determined variation in the PSC chemical phenotype (hereafter referred to as chemotype) is present in many species. For example, a strong genetic basis for variation in the identity of leaf chemical compounds has been demonstrated for terpenes produced in oak and pine trees [29] [30, 31], aromatic Lamiaceae [32], Asteraceae [33], for glucosinolates produced in the Brassicaceae [34], and for flavonoids in Berberidaceae [35], illustrating that genetic variation for these compounds are present in a wide range of plant families. Thus the (a)biotic environment can shape which compounds are released to the environments not only through differential induction of PSCs, but also by natural selection, favouring genotypes producing specific chemotypes (Figure 1)[36]. Differences in climatic conditions and herbivore pressure favour specific chemotypes both within and among species [29, 37, 38]. A recent example of evolutionary changes in PSC production in response to climate change

is that of wild thyme (Thymus vulgaris), a small shrub widespread in the Mediterranean Basin. The dominant monoterpene in thyme is genetically determined to be a phenolic or a non-phenolic type. These two chemotypes are ecotypes with adaptation to warm dry summers and early winter frost, respectively [39]. The lack of very cold winters in the last 25 years is associated with a significant increase in phenolic chemotypes in regions that were previously dominated by non-phenolic types [38]. In addition to differences in their ability to tolerate frost and drought, the toxicities of phenolic and non-phenolic chemotypes differ towards associated plants, microorganisms and herbivores, suggesting that climate-driven changes in the genetic composition of this widespread species may have cascading effects on associated species interactions.

A correlation between variation in PSC production and environmental variation holds for many plant species, but we need more long-term experimental studies to understand whether current and future changes in climate and herbivore pressure will result in major induced or genetic changes in PSC production. If so, cascading effects of such changes are expected for belowground species interactions and decomposition of organic matter and ultimately nutrient cycling in soil [25, 40]. Moreover, climate change may also affect the concentration of PSCs in soil by altering soil-water content, temperature, and pH. As a result, the levels and circulation of liquid and gas among soil pores (see below) will change and hence the movement and diffusion of PSCs.

Diffusion of PSC in soil

In contrast to atmospheric PSCs, volatile PSCs in the soil can form stable concentration gradients due to the lack of air turbulence. The diffusion of PSCs in soil depends on their physicochemical properties (Table 1) and the texture of the soil matrix. Fine-textured and clayey soils facilitate the flow of volatile PSCs that are diluted in smaller air volumes due to the smaller pores of these soil types. This generates gradients that are steeper and propagate faster than those in soils with large air spaces. In well-ventilated soils, though, PSCs will rapidly fill large air volumes, albeit at lower concentrations.

PSC transfer in soil is also affected by adsorption to the mineral and organic surfaces of soil particles, whose fractions depend on soil type and layer, with A horizons (topsoil) generally containing more organic material than the underlying B horizons (subsoil) [41, 42]. PSCs will adhere to and dissolve on lipo- and hydrophilic surfaces during diffusion, thus forming intermediate pools in equilibrium

with the gas-phase concentrations [43, 44]. Soil organisms, therefore, receive PSCs more slowly and with an attenuated amplitude compared to the actual variations in the emission strength of the PSC source.

Abiotic factors such as temperature, soil moisture, and pH interfere with the persistence and diffusion of PSCs in soil. The adsorption of volatile PSCs by the mineral and organic fractions of soil depends on pH, with higher adsorption in alkaline than acidic soils [45]. The acidity of soil-pore water can affect the uptake of volatiles on aqueous surfaces, e.g. by protonating highly reactive sesquiterpenes that are otherwise poorly soluble in water [46]. Periods of rain and drought change the movement patterns of PSCs through soil, with opposite behaviours expected for volatile and non-volatile water-soluble PSCs (Table 1). During rainy weather, volatile PSCs are first concentrated in the soil pores during water penetration and then move with air bubbles to the surface [47], whereas non-volatile water-soluble PSCs are diluted and transported away from their sources [48]. High levels of soil moisture also affect the availability of oxygen for roots and soil microorganisms, eventually generating anaerobic conditions, which influence both plant PSC-producing and microbial PSC-decomposing metabolisms [45, 49]. In contrast, drought may decrease the concentration of volatile PSCs in the soil due to the increase in the volume of air in the soil and better ventilation at the soil surface. Sesquiterpenes, for example, emitted by maize diffuse faster and farther at low moisture levels [50].

Climate change can deeply impact the movement and lifetime of PSCs in soils, either by affecting abiotic drivers directly or by affecting the macro- and micro-biotic communities involved in the liberation, processing and transport of PSCs [51] (see also section "Plant-soil microorganism interactions" below). For example, increasing soil temperature will enhance PSC diffusion and exchange rates and alter state and partitioning equilibria between surfaces. Extreme weather events such as thunderstorms and drought events are expected to increase in both frequency and intensity, which will amplify the asymmetric transport and concentration build-up of polar/non-polar PSCs in soils. Determining the presence, and the fate of PSCs in soil is an under-researched area but is highly relevant to evaluate their ecological and evolutionary impact on the belowground interactions discussed below. Understanding the main environmental factors driving these processes is necessary to assess how climate change will alter belowground interactions mediated by PSCs (Figure 1). Different methods may be needed for different types of PSCs, depending on their solubility and volatility (see Box 1, Table 1).

Ecological and evolutionary impacts of PSCs on belowground species interactions

192

194

195

196

191

193 Similar to the functions of PSCs aboveground, key organismal interactions in soil, i.e. plant-plant,

plant-soil microorganism, and plant-soil invertebrate interactions, are also mediated by PSCs. The

identity, concentration, and composition of PSCs belowground play pivotal roles in species

interactions and community structuring, and ultimately affect entire ecosystem processes.

197

198

Plant-plant interactions

- 199 PSCs mediate both competitive and facilitative interactions among plants. Allelopathic effects of
- 200 PSCs released by a focal plant to the local soil environment can inhibit germination and growth of
- 201 competitors giving the focal plant a competitive advantage for space and nutrients [52]. The "Novel
- Weapons" hypothesis [53] is based on the idea that invasive plants release allelochemicals which the
- 203 native community has not previously encountered and that give the invader a competitive advantage.
- 204 Allelopathic effects of PSC have helped some invasive plant species to create virtual monocultures
- in the habitats they invade. For example, production of **phytotoxic** phenolic acids, and di-, and
- sesquiterpene lactones likely aided the invasive success of Canadian goldenrod Solidago canadiensis
- 207 [54, 55] and yellow star thistle *Centaurea solstitialis* [56]. Allelopathic effects are often less dramatic
- 208 within the native ranges of plants because coexisting species have evolved a tolerance to local PSCs,
- or the ability to detoxify them [57].
- 210 PSCs released to the soil by one plant species may alter competitive hierarchies among neighbouring
- 211 plant species both directly when PSC inhibit growth of one species more than another, and indirectly
- by PSC modifying the soil nutrient availability that favour some species over others. For example,
- 213 pine soil altered plant competition between spotted knapweed (Centaurea stoebe) and different grass
- species in favour of grasses. Soil containing pine litter had a higher content of phosphorus (P), and
- 215 the higher availability of P made the grasses more competitive against Centaurea [58]. Phenolic
- 216 compounds and terpenes influence soil nitrogen (N) cycling as these compounds inhibit nitrification
- and the net N mineralization [59]. This could differentially favour plant species that differ in their
- 218 preference for nitrate vs ammonium [60].
- 219 PSC mediated plant-plant interactions also facilitate neighbour plants via so-called associational
- resistance [61] where neighbour plants benefit from reduced attack by herbivores and pathogens due

- 221 to the deterrent effects of the PSCs released by a focal plant. However, the PSCs in soil may also
- 222 increase susceptibility to herbivores and pathogens if these plant enemies are attracted rather than
- repelled by the PSC [52, 61]. How these selective forces are balanced in the multispecies networks
- present in a natural field needs further study.
- 225 PSCs can vary among species, genotypes within species, and damaged and healthy tissue, so they
- 226 contain information about the identity and state of the emitting plant [62]. Plants may therefore use
- PSCs as a cue to recognize their neighbours and prepare for the competitive encounters by adjusting
- 228 their competitive growth towards or away from the neighbouring plant [63]. Root exudates and the
- chemicals they contain have been shown to mediate recognition of neighbour plants [64-66]. The
- competitive responses to such PSC-mediated neighbour recognition may depend on the history of co-
- existence between interacting plants [67], and for some plant species also on the genetic relatedness
- among interacting conspecifics [64, 68]. However, we lack knowledge on the identity of which
- 233 molecules in root exudates that mediate plant recognition and how plants perceive them. This
- knowledge is relevant in order to understand if environmental changes in the soil may interfere with
- 235 plants ability to detect their neighbour.
- Spatial consistency of PSC production allows PSCs to act as selective agents on other species and
- 237 favour the genotypes that respond most beneficially to the effects of frequently encountered PSCs.
- For example, the grass *Bromus erectus* is a frequent neighbour of *T. vulgaris*, which releases either
- phenolic or non-phenolic monoterpenes to the local soil environment. Bromus plants originating from
- sites where their thyme neighbours produce a non-phenolic monoterpene germinate and grow better
- in soil rich in non-phenolic thyme monoterpenes compared to similar soil rich in phenolic thyme
- 242 monoterpenes [69]. Hence, plants can adapt to the chemical environment imposed by frequent
- 243 neighbours, and this adaptation is highly compound-specific. Changes in the consistency of
- 244 frequently encountered PSCs could disrupt such adaptive responses.
- 245 From pairwise plant-plant interaction to communities
- 246 The PSCs released to the soil from different plant genotypes and species create a mosaic of soil
- 247 chemical environments. Because different chemical environments favour different species due to the
- 248 pairwise competitive and facilitative interactions described above, variation in chemical types both
- 249 within and among populations can contribute to local and regional species richness. For example, the
- 250 chemical diversity of Scots Pine (Pinus sylvestris) needle terpenes was positively correlated with
- 251 plant species richness in the vegetation under the trees, and trees with different needle chemistry

associated with a different plant species composition [70, 71]. Chemical diversity of *T. vulgaris* also correlated with plant species richness and composition of the vegetation around thyme plants. Plant species richness was higher, and species composition more similar around thyme plants, than in thyme-free area only few meters away. However, species richness and species composition differed in area with the different thyme chemical types [72]. These examples suggest that the chemical types of dominant plants have an "extended phenotype" [73, 74] where the chemicals released from a focal plant can shape the surrounding vegetation. However, we need more experimental manipulation studies investigating the consistency of the effects of variation in PSC released by foundation plants on richness and composition of associated plant communities. These studies will allow us to better understand cause and effect of variation in PSC and to predict the potential cascading effects of qualitative and quantitative changes in PSCs.

Plant-soil microorganism interactions

A large proportion of the nutrients present in the soil are not immediately available for plants but require enzymes to degrade compounds into usable parts, and soil microorganisms are major suppliers of these enzymes [80]. In addition, soil symbionts such as mycorrhizal fungi and rhizobacteria are important regulators of plant productivity especially in nutrient poor ecosystems, where they are responsible for 5 to 75% of nitrogen and up to 80% of phosphorous acquired by plants annually [81] The importance and complexity of the interactions between plants and soil microbes have been acknowledged for a long time, but only more recently it has been appreciated that many of these interactions are mediated by PSCs [82]. Given the importance of such interactions, it is likely that plants have evolved strategies to recognize friends from foes and to harness useful partners across distances. For example, flavonoids produced by legumes have a fundamental role in attracting rhizobia and promoting the production of Nod-factor molecules, which are critical for nodulation [83]. In this process, both plants and microbes produce chemical signals that allow the plant to recognize the rhizobia as a friend and prevent the activation of immune responses upon initiation of the symbiosis [84]. When rhizobia are not found in the proximity of the rhizosphere and distances between the roots of individual plants and their symbionts are too large to be covered by root exudates, legumes can resort to the help of an intermediary to carry the symbionts to them [83, 85]. The production of volatile PSCs by the legume Medicago truncatula attracts individuals of the freeliving nematode Caenorhabditis elegans, which harbours the rhizobium Sinorhizobium meliloti either

on its cuticle or in its intestine. Isoflavonoids produced by the plant prime the rhizobium to initiate the symbiosis when the nematode is near the roots [75].

Not only can plants affect the composition of specific soil microbial communities via PSC, but they can also alter it when exposed to stress such as herbivore attacks [82]. In many plant species, the presence of pathogens or root feedings fungi corresponds with a shift in the soil microbial composition close to the rhizosphere. This is a consequence of both the production of specific antimicrobial compounds by the plant rhizosphere and of recruitment of favourable bacteria [86] The blend of PSCs produced by the graminoid *Carex arenaria* shifts toward a higher concentration of terpenes and terpenoids [59] to attract specific bacteria with antifungal properties when the roots are infected with the pathogenic fungus *Fusarium culmorum* [87]. Bacteria successfully rallied by the plant can gain from this interaction via the additional nutrients obtained by feeding on the plant pathogens [87]. Recent studies compared wild types of *Arabidopsis thaliana* with mutants that are unable to synthesize PSCs involved in induced systemic resistance against soil pathogens. Metagenome analysis of root microbiome associated with wild types vs. mutants showed substantially different community composition, indicating the important role of PSC in modulating the composition of root microbiome [76].

- PSC-induced shifts in the composition of soil microbiota can also be activated by attacks to plant leaves. For example, infestations of white flies (*Bemisia tabaci*) in pepper plants lead to a change in belowground microflora enhancing anti-pathogen responses [88].
- The interactions between plants and soil microorganisms are not unidirectional, as soil microbes can also emit chemical compounds that can be detected and used by plants. For example, secondary compound emissions from soil microflora may help the woodland strawberry, *Fragaria vesca*, to detect nutrient-rich patches prior to rooting. Stolons can preferentially develop toward substrates emitting unbranched fatty acids and avoid substrates with higher concentrations of terpenes and aldehydes [89].

Many PSCs are used by bacteria and fungi as a source of nutrients [90, 91], and soil microorganisms can alter and repurpose the PSCs released by plants. In the legume *Lotus japonicus*, the inoculation with its symbiont *Mesorhizobium loti* leads to a change in the profile of the plant phenolic compounds [60]. The soil bacterial community associated with vetiver, *Chrysopogon zizanioides*, uses the sesquiterpenes produced by the grass as a source of carbon, and after metabolizing them releases several compounds from vetiver oil [63]. Colonization of roots by arbuscular mycorrhizal fungi can

increase the production of glucosinolates, phenolic acids, and flavonoids in several plant species [64, 65].

On a large scale, the ability of soil microbial communities to process and degrade plant secondary

metabolites can have important consequences on plant communities and ecosystem stability. For

example, soil microorganisms can either favour or hinder the proliferation on invasive plant species,

depending on their ability to degrade the allelopathic compounds produced by invasive plants [92].

321 A high species-specificity of plant-microbe interactions suggest that single plants can shape their soil

bacterial communities, and microorganisms can respond to subtle genetic variation in plants. In

cottonwood trees Populus angustifolia, 70% of the variation in soil microbial community

composition was explained by plant genotype identity [77].

Although there is still much that we need to understand about the complex interactions between PSC and roots microbiome, progress is being made in elucidating the factors that drive the composition of root inhabiting bacterial communities [78, 79]. Future steps will need to go beyond pairwise interactions between plants and a specific guild of soil microorganism, targeting instead how whole communities of soil microorganisms are affected by PSC and in turn affect plant communities. Climate change may mediate shifts in plant-soil microorganism interactions via shifts in the chemical profile of plants and soil microorganisms. Such studies are highly valuable as the positive effect of certain plant-soil microorganism interactions can be harnessed to either enhance agricultural practices or as biocontrol agents. Understanding how to engineer beneficial soil communities for plant community conservation, sustainable agriculture and preservation of native communities require extensive experimentation and in depth understanding of potential unintended consequences of such manipulations.

Plant-soil invertebrate interactions

Many of the interactions between plants and soil invertebrates are mediated by PSCs, often with significant effects on soil processes that determine soil function and plant productivity. Two major ways can be distinguished through which PSCs mediate the interactions between plants and soil invertebrates. First, soil organisms can be exposed to substantial input of PSCs from aboveground plant tissues, through PSCs leaching from foliar tissue and PSC-rich litter deposition. These PSCs are mostly known as feeding repellents, and many of them with toxic effects to aboveground insect

herbivores [80], and the available evidence indicates that they also have toxic effects on non-target organisms such as soil invertebrates [81, 82]. Toxic effects can occur if PSCs diffuse across the cuticle or if animals ingest PSC-rich litter, with the main route of exposure determined by cuticle morphology and diet of the soil invertebrate and by the chemical properties of the PSC (see Table 1 for mode of toxicity). Nonpolar (lipophilic) volatile PSCs likely enter animal bodies primarily by simple diffusion across the cuticle and easily partition into cellular membranes [81]. High concentrations of nonpolar PSCs can occur regularly in soil, especially in older litter and near roots, as is the case for terpenoids (1-5 mg g⁻¹ fresh weight) [2]. The likelihood of exposure is increased because terpenoids will remain in the litter layer for several months until fully decomposed [83]. Exposure to polar PSCs is mainly through ingestion of litter or roots, for example in case of isothiocyanates. Isothiocyanates have a high toxicity towards non-target soil invertebrates, including beneficial macro-detritivores that promote nutrient cycling and soil fertility [84]. Currently, the relative importance of the two exposure routes, either via contact with volatile or water dissolved PSCs in pore water and diffusion across the cuticula or via ingestion of PSC-rich litters to soil fauna and how this affects soil fauna functional community composition is poorly known.

An unexplored aspect of the lipophilic PSCs is that their spread can be facilitated by animal transport. Soil animals can accumulate high internal concentrations of these PSCs in the vicinity of roots and litter, and subsequently, deliver them to other soil microenvironments with lower chemical activity when they move away from the PSC source. Aquatic Protozoa have demonstrated such animal-facilitated transport [85], but more research is needed to explore this phenomenon mediated by soil animals. In order to explore the toxicity of PSCs to soil invertebrates under field conditions, more precise analytical determinations of concentrations of PSCs and their degradation products in the environment are needed. This necessitates new experimental systems of exposure, e.g. by using the framework of chemical activity for volatile compounds [81].

The second way in which PSCs can mediate plant invertebrate interactions in the soil is by targeted release of PSCs from the roots. These PSCs can attract or deter soil invertebrates and serve as chemical signals released by the plant to control trophic interactions with plant-parasitic nematodes, root herbivores, and their predators. The use of PSCs in tri-trophic interactions between plants and belowground organisms may be similar to the extensively studied chemically mediated interactions between plants, foliar herbivores, and their natural aboveground predators. Attack by root herbivores has been shown to induce a change in the quality and quantity of PSCs released by plant roots, which can then function as a signal to attract soil-dwelling predators and reduce the pest population [9, 86,

87]. Entomopathogenic nematodes often function as indirect defences for plants [88, 89]. For example, larvae of the longhorn beetle Tetraopsis tetraophthalmus feed on the roots of the common milkweed Asclepias syriaca which releases volatiles into the soil that attract entomopathogenic nematodes, preventing the loss of plant biomass to herbivory [90]. Maize roots increase sesquiterpene exudation when attacked by insects, thus attracting entomopathogenic nematodes that feed on root herbivores [91]. However, notwithstanding this potential protective role of PSCs via attraction of beneficial nematodes, these PSCs can also attract the root herbivores themselves. The net effect of the attraction of both advantageous as well as detrimental nematodes on plant fitness has to be evaluated. Most examples of tri-trophic interactions belowground involve entomopathogenic nematodes, but one of the few papers that reports on other soil fauna in tri-trophic interactions is a case in which tulip bulbs infested by the rust mite Aceria tulipae, emits volatile signals to the predatory mite Neoseiulus cucumeris [92]. Given the importance of chemical cues in belowground signalling, the effect of PSCs on tri-trophic interactions is expected to be omnipresent, but currently, we do not know enough about the spatial dimensions of volatile and water-dissolved PSCs effects on plant-root herbivore or parasite-herbivore interactions [93]. Abiotic gradients in soils are rather short and steep compared to aboveground systems, but currently, we are ignorant of the spatial dimensions of volatile and water-dissolved PSCs effects on plant-root herbivore or parasite-predator interactions.

PSCs and belowground co-evolutionary interactions

The examples above demonstrate the different vital roles that PSCs play in belowground plant-plant, plant-soil microorganism, and plant-soil invertebrate interactions. Some of these interactions are examples of the evolutionary adaptation of species to frequently encountered PSCs. Except for interactions specific to the rhizosphere, no studies have yet clearly shown whether adaptations of organisms to specific PSCs reciprocally alter PSC composition and production in plants, i.e. if PSC-mediated belowground interactions result in pairwise or diffuse co-evolution. Understanding if belowground interactions are a source of selection on the chemical variation of PSC-producing plants is not straightforward because of the unspecific nature of the interactions. Belowground interactions often consist of multiple interactions, therefore, estimating the consequences for fitness in PSC-producing plants is difficult. Also, for PSCs entering the soil via leaf leachates, the main selective force determining which PSCs are produced and in what concentration may largely depend on aboveground biotic interactions and local climate. In addition, the multifunctionality of PSCs extends

beyond biotic interactions, which is a further complication in the evaluation of fitness consequences in PSC-producing plants. For instance, maize produces benzoxazinoid compounds not only as a defence against generalist herbivores but also as chelating agents to acquire iron as a micronutrient. The larvae of western corn rootworm *Diabrotica virgifera*, are not injured by benzoxazinoids, but this specialist herbivore instead takes advantage of the presence of complexes between iron and these PSCs not only to detect the plant but also forage on these complexes [94].

Climate-induced shifts in the release of PSCs into the soil (Figure 1) may alter the outcome of within-and between-species interactions in ways that we still cannot predict. Long-term studies and the development of ecometabolomics [95] for elucidating the responses of PSC changes in plant communities due to warming and altered precipitation from climate change are particularly warranted for a better understanding of, and ability to predict, these shifts in PSCs production, diffusion in soil, and effects on belowground interactions.

420

421

422

423

424

425

426

427

428

429

430

431

432

433

434

435

436

437

438

414

415

416

417

418

419

Concluding remarks and perspectives

Soil is a theatre of facilitation, symbiosis and warfare deployed by plants and the various organisms living in it, and PSCs plays a major role mediating many of these interactions. Plants and soil organisms have adapted to withstand, detoxify or use the cocktail of PSCs originally meant to harm some of them. Understanding PSC mediated relationships at community scale and identifying the compounds involved in these interactions is therefore important for better insight into the functioning of these systems and their evolution, especially in changing environments. High-throughput technologies are increasingly becoming available and may help to identify and determine the concentrations, transport and modes of action of PSCs in soils. We encourage the creation of open, community-wide, curated, labelled, broad-spectrum PSC data sets across plant species and soils, as this would greatly increase the transfer of knowledge between scientist studying plants, microbes, and invertebrates in this biological belowground theatre. This could be a platform similar to, e.g. the TRY plant trait data base (http://www.try-db.org) or PSC data sets could be added and incorporated to such established well-functioning platforms. The warranted study of the quantitative and qualitative effects of climate change on PSCs and their interactions with the environment and organisms would also benefit from these PSC data sets. The alterations of the production of PSCs due to climate change are highly variable, depending on dose, timing, PSC, and species. These changes in the production of PSCs can lead to unforeseen consequences for soil structure and function

- and can disturb biological feedbacks on soil chemistry and biology, perhaps even on atmospheric
- chemistry and climate [96], with a direction and intensity that warrants in-depth investigation.

References

442

441

- 1. Moore, B.D. et al. (2014) Explaining intraspecific diversity in plant secondary metabolites in an
- ecological context. New Phytol. 201, 733-750.
- 2. Alessio, G.A. et al. (2008) Implications of foliar terpene content and hydration on leaf
- flammability of *Quercus ilex* and *Pinus halepensis*. *Plant Biol.* 10, 123-128.
- 3. Gershenzon, J. and Dudareva, N. (2007) The function of terpene natural products in the natural
- 448 world. Nat. Chem. Biol. 3, 408-414.
- 4. Baldwin, I.T. et al. (2006) Volatile signaling in plant-plant interactions:" talking trees" in the
- 450 genomics era. *Science* 311, 812-815.
- 451 5. Hare, J.D. (2011) Ecological role of volatiles produced by plants in response to damage by
- 452 herbivorous insects. Annu. Rev. Entomol. 56, 161-180.
- 453 6. Ninkovic, V. et al. (2016) Decoding neighbour volatiles in preparation for future competition and
- 454 implications for tritrophic interactions. *Perspect. Plant Ecol. Evol. Syst.* 23, 11-17.
- 7. Delory, B.M. et al. (2016) Root-emitted volatile organic compounds: can they mediate
- belowground plant-plant interactions? Plant Soil 402, 1-26.
- 8. Sasse, J. et al. (2018) Feed your friends: do plant exudates shape the root microbiome? Trends
- 458 Plant Sci. 23, 25-41.
- 9. Rasmann, S. and Turlings, T.C. (2016) Root signals that mediate mutualistic interactions in the
- 460 rhizosphere. Curr. Opin. Plant Biol. 32, 62-68.
- 461 10. Massalha, H. et al. (2017) Small molecules below- ground: the role of specialized metabolites
- 462 in the rhizosphere. *Plant J.* 90, 788-807.
- 463 11. Maruri-López, I. et al. (2019) Intra and extracellular journey of the phytohormone salicylic acid.
- 464 Front. Plant Sci. 10, Article 423.
- 465 12. Park, J.H. et al. (2013) Active atmosphere-ecosystem exchange of the vast majority of detected
- volatile organic compounds. Science 341, 643-647.
- 13. Nguyen, T.B. et al. (2015) Rapid deposition of oxidized biogenic compounds to a temperate
- 468 forest. *Proc. Natl. Acad. Sci.* 112, E392-E401.

469 14. Spielmann, F.M. et al. (2017) Isoprene and α-pinene deposition to grassland mesocosms. Plant

- 470 Soil 410, 313-322.
- 471 15. Potier, E. et al. (2017) Chemical reaction rates of ozone in water infusions of wheat, beech, oak
- and pine leaves of different ages. Atmos. Environ. 151, 176-187.
- 473 16. Gray, C.M. et al. (2010) Emissions of volatile organic compounds during the decomposition of
- plant litter. J. Geophys. Res. Biogeo. 115, G03015.
- 475 17. Omirou, M. et al. (2013) Dissipation of pure and broccoli-released glucosinolates in soil under
- high and low moisture content. Eur. J. Soil Biol. 56, 49-55.
- 18. Staudt, M. et al. (2019) Compartment specific chiral pinene emissions identified in a Maritime
- 478 pine forest. Sci. Total Environ. 654, 1158-1166.
- 19. Mäki, M. et al. (2019) Boreal forest soil is a significant and diverse source of volatile organic
- 480 compounds. Plant Soil, 1-22.
- 481 20. Gimsing, A.L. and Kirkegaard, J.A. (2009) Glucosinolates and biofumigation: fate of
- 482 glucosinolates and their hydrolysis products in soil. *Phytochem. Rev.* 8, 299-310.
- 483 21. Gimsing, A.L. et al. (2005) Extraction and determination of glucosinolates from soil. J. Agric.
- 484 *Food Chem.* 53, 9663-9667.
- 485 22. Cantor, A. et al. (2011) Low allelochemical concentrations detected in garlic mustard-invaded
- forest soils inhibit fungal growth and AMF spore germination. *Biol. Invasions* 13, 3015-3025.

- 487 23. Possell, M. and Loreto, F. (2013) The role of volatile organic compounds in plant resistance to
- 488 abiotic stresses: responses and mechanisms. In Biology, controls and models of tree volatile organic
- compound emissions (Niinemets, U. and Monson, R.K. eds), pp. 209-235, Springer Netherlands.
- 490 24. Sobuj, N. et al. (2018) Impacts of elevated temperature and CO₂ concentration on growth and
- 491 phenolics in the sexually dimorphic *Populus tremula* (L.). *Environ. Exp. Bot.* 146, 34-44.
- 492 25. Holopainen, J.K. et al. (2018) Climate change effects on secondary compounds of forest trees in
- 493 the northern hemisphere. Front. Plant Sci. 9, 1445.
- 494 26. Peñuelas, J. and Estiarte, M. (1998) Can elevated CO₂ affect secondary metabolism and
- 495 ecosystem function? Trends Ecol. Evol. 13, 20-24.
- 496 27. Haberstroh, S. et al. (2018) Terpenoid emissions of two Mediterranean woody species in
- 497 response to drought stress. Front. Plant Sci. 9, 1071.
- 498 28. Bonn, B. et al. (2019) Biogenic isoprenoid emissions under drought stress: different responses
- 499 for isoprene and terpenes. *Biogeosciences* 16, 4627-4645.
- 500 29. Iason, G.R. et al. (2012) The ecology of plant secondary metabolites: from genes to global
- 501 processes, Cambridge University Press.
- 30. Song, W. et al. (2014) Laboratory and field measurements of enantiomeric monoterpene
- emissions as a function of chemotype, light and temperature. *Biogeosciences* 11, 1435-1447.
- 31. Staudt, M. et al. (2004) Isoprenoid emissions of Quercus spp.(Q. suber and Q. ilex) in mixed
- stands contrasting in interspecific genetic introgression. New Phytol. 163, 573-584.
- 32. Thompson, J.D. (2005) *Plant evolution in the Mediterranean*, Oxford University Press on
- 507 Demand.
- 508 33. Pratt, J.D. et al. (2014) Genetically based latitudinal variation in Artemisia californica
- secondary chemistry. Oikos 123, 953-963.
- 34. Brachi, B. et al. (2015) Coselected genes determine adaptive variation in herbivore resistance
- 511 throughout the native range of Arabidopsis thaliana. Proc. Natl. Acad. Sci. 112, 4032-4037.
- 512 35. Chen, J. et al. (2015) Chemotypic and genetic diversity in Epimedium sagittatum from different
- 513 geographical regions of China. *Phytochemistry* 116, 180-187.
- 36. Fischer, D. et al. (2014) Plant genetic effects on soils under climate change. Plant Soil 379, 1-
- 515 19.
- 37. Poelman, E.H. and Kessler, A. (2016) Keystone herbivores and the evolution of plant defenses.
- 517 *Trends Plant Sci.* 21, 477-485.
- 38. Thompson, J. et al. (2013) Evolution of a genetic polymorphism with climate change in a
- 519 Mediterranean landscape. Proc. Natl. Acad. Sci. 110, 2893-2897.
- 39. Thompson, J.D. et al. (2007) Ongoing adaptation to Mediterranean climate extremes in a
- 521 chemically polymorphic plant. Ecol. Monogr. 77, 421-439.
- 522 40. Adamczyk, B. et al. (2018) Plant Secondary Metabolites—Missing Pieces in the Soil Organic
- 523 Matter Puzzle of Boreal Forests. *Soil Systems* 2, 2.
- 41. Kanerva, S. et al. (2008) Phenolic compounds and terpenes in soil organic horizon layers under
- silver birch, Norway spruce and Scots pine. *Biol. Fertility Soils* 44, 547-556.
- 42. Uteau, D. et al. (2013) Root and time dependent soil structure formation and its influence on gas
- transport in the subsoil. *Soil and Tillage Research* 132, 69-76.
- 43. Provoost, J. et al. (2011) Henry's equilibrium partitioning between ground water and soil air:
- Predictions versus observations. J. Environ. Prot. (Irvine, Calif.) 2, 873-881.
- 530 44. Hamamoto, S. et al. (2012) Organic matter fraction dependent model for predicting the gas
- diffusion coefficient in variably saturated soils. Vadose Zone J. 11, 1539-1663.
- 45. Insam, H. and Seewald, M.S. (2010) Volatile organic compounds (VOCs) in soils. *Biol.*
- 533 *Fertility Soils* 46, 199-213.

- 46. Matsuoka, K. et al. (2017) Reactive uptake of gaseous sesquiterpenes on aqueous surfaces. J.
- 535 *Phys. Chem. A* 121, 810-818.
- 47. Bourtsoukidis, E. et al. (2018) Strong sesquiterpene emissions from Amazonian soils. Nat.
- 537 *Commun.* 9, 2226.
- 48. Clauson-Kaas, F. et al. (2016) Ptaquiloside from bracken in stream water at base flow and
- during storm events. Water Res. 106, 155-162.
- 540 49. Seewald, M.S. et al. (2010) Substrate-induced volatile organic compound emissions from
- 541 compost-amended soils. Biol. Fertility Soils 46, 371-382.
- 50. Hiltpold, I. and Turlings, T.C. (2008) Belowground chemical signaling in maize: when
- simplicity rhymes with efficiency. J. Chem. Ecol. 34, 628-635.
- 51. Glassman, S.I. et al. (2018) Decomposition responses to climate depend on microbial
- community composition. *Proc. Natl. Acad. Sci.* 115, 11994-11999.
- 52. Effah, E. et al. (2019) Potential roles of volatile organic compounds in plant competition.
- 547 Perspect. Plant Ecol. Evol. Syst. 38, 58-63.
- 53. Callaway, R.M. and Ridenour, W.M. (2004) Novel weapons: invasive success and the evolution
- of increased competitive ability. Front. Ecol. Environ. 2, 436-443.
- 550 54. Yuan, Y. et al. (2013) Enhanced allelopathy and competitive ability of invasive plant Solidago
- 551 canadensis in its introduced range. J. Plant Ecol. 6, 253-263.
- 55. Abhilasha, D. et al. (2008) Do allelopathic compounds in invasive Solidago canadensis sl
- restrain the native European flora? J. Ecol. 96, 993-1001.
- 56. Irimia, R.E. et al. (2019) Biogeographic differences in the allelopathy of leaf surface extracts of
- an invasive weed. Biol. Invasions 21, 3151-3168.
- 57. Inderjit et al. (2011) The ecosystem and evolutionary contexts of allelopathy. Trends Ecol. Evol.
- 557 26, 655-662.
- 58. Metlen, K.L. and Callaway, R.M. (2015) Native North American pine attenuates the
- competitive effects of a European invader on native grasses. Biol. Invasions 17, 1227-1237.
- 560 59. Smolander, A. et al. (2012) Nitrogen transformations in boreal forest soils—does composition
- of plant secondary compounds give any explanations? *Plant Soil* 350, 1-26.
- 562 60. Boudsocq, S. et al. (2012) Plant preference for ammonium versus nitrate: a neglected
- determinant of ecosystem functioning? Am. Nat. 180, 60-69.
- 61. Barbosa, P. et al. (2009) Associational resistance and associational susceptibility: having right
- or wrong neighbors. Annu. Rev. Ecol. Evol. Syst. 40, 1-20.
- 62. Ninkovic, V. et al. (2019) Who is my neighbor? Volatile cues in plant interactions. Plant
- 567 Signal. Behav. 14, e1634993.
- 568 63. Novoplansky, A., What plant roots know?, Seminars in cell & developmental biology, Elsevier,
- 569 2019.
- 570 64. Semchenko, M. et al. (2014) Plant root exudates mediate neighbour recognition and trigger
- 571 complex behavioural changes. New Phytol. 204, 631-637.
- 572 65. van Dam, N.M. and Bouwmeester, H.J. (2016) Metabolomics in the rhizosphere: tapping into
- belowground chemical communication. *Trends Plant Sci.* 21, 256-265.
- 66. Kong, C.-H. et al. (2018) Plant neighbor detection and allelochemical response are driven by
- 575 root-secreted signaling chemicals. *Nat. Commun.* 9, 1-9.
- 576 67. Semchenko, M. et al. (2019) Soil biota and chemical interactions promote co- existence in co-
- evolved grassland communities. *J. Ecol.* 107, 2611-2622.
- 68. Biedrzycki, M.L. et al. (2010) Root exudates mediate kin recognition in plants. Commun.
- 579 *Integr. Biol.* 3, 28-35.

- 580 69. Ehlers, B.K. and Thompson, J. (2004) Do co-occurring plant species adapt to one another? The
- response of Bromus erectus to the presence of different *Thymus vulgaris* chemotypes. *Oecologia*
- 582 141, 511-518.
- 583 70. Iason, G.R. et al. (2005) Does chemical composition of individual Scots pine trees determine
- the biodiversity of their associated ground vegetation? *Ecol. Lett.* 8, 364-369.
- 585 71. Pakeman, J.R. et al. (2006) The extended phenotype of Scots pine Pinus sylvestris structures the
- understorey assemblage. *Ecography* 29, 451-457.
- 587 72. Ehlers, B.K. et al. (2014) An allelopathic plant facilitates species richness in the Mediterranean
- 588 garrigue. *J. Ecol.* 102, 176-185.
- 589 73. Dawkins, R. (1982) The extended phenotype, Oxford University Press Oxford.
- 590 74. Whitham, T.G. *et al.* (2003) Community and ecosystem genetics: a consequence of the extended phenotype. *Ecology* 84, 559-573.
- 75. Horiuchi, J.-i. *et al.* (2005) Soil nematodes mediate positive interactions between legume plants and rhizobium bacteria. *Planta* 222, 848-857.
- 76. Stringlis, I.A. et al. (2018) MYB72-dependent coumarin exudation shapes root microbiome
- assembly to promote plant health. Proc. Natl. Acad. Sci. 115, E5213-E5222.
- 596 77. Schweitzer, J.A. et al. (2008) Plant–soil–microorganism interactions: heritable relationship
- between plant genotype and associated soil microorganisms. *Ecology* 89, 773-781.
- 598 78. Bulgarelli, D. et al. (2012) Revealing structure and assembly cues for Arabidopsis root-
- inhabiting bacterial microbiota. *Nature* 488, 91-95.
- 600 79. Walters, W.A. et al. (2018) Large-scale replicated field study of maize rhizosphere identifies
- heritable microbes. *Proc. Natl. Acad. Sci.* 115, 7368-7373.
- 80. Mithöfer, A. and Boland, W. (2012) Plant defense against herbivores: chemical aspects. Annu.
- 603 Rev. Plant Biol. 63, 431-450.
- 81. Trac, L. et al. (2019) Headspace passive dosing of volatile hydrophobic organic chemicals from
- a lipid donor linking their toxicity to well-defined exposure for an improved risk assessment.
- 606 Environ. Sci. Technol. 53, 13468-13476.
- 82. van Ommen Kloeke, A.E. et al. (2012) Time-related survival effects of two gluconasturtiin
- 608 hydrolysis products on the terrestrial isopod *Porcellio scaber*. Chemosphere 89, 1084-1090.
- 83. Kainulainen, P. and Holopainen, J. (2002) Concentrations of secondary compounds in Scots
- pine needles at different stages of decomposition. Soil Biol. Biochem. 34, 37-42.
- 84. Jensen, J. et al. (2010) The toxic effects of benzyl glucosinolate and its hydrolysis product, the
- biofumigant benzyl isothiocyanate, to Folsomia fimetaria, Environ. Toxicol. Chem. 29, 359-364.
- 85. Gilbert, D. et al. (2014) Co-transport of polycyclic aromatic hydrocarbons by motile
- 614 microorganisms leads to enhanced mass transfer under diffusive conditions. Environ. Sci. Technol.
- 615 48, 4368-4375.
- 86. Wenke, K. et al. (2010) Belowground volatiles facilitate interactions between plant roots and
- 617 soil organisms. *Planta* 231, 499-506.
- 87. Kergunteuil, A. et al. (2019) Environmental gradients and the evolution of tri- trophic
- 619 interactions. *Ecol. Lett.* 22, 292-301.
- 88. Laznik, Ž. and Trdan, S. (2016) Attraction behaviors of entomopathogenic nematodes
- 621 (Steinernematidae and Heterorhabditidae) to synthetic volatiles emitted by insect damaged potato
- 622 tubers. *J. Chem. Ecol.*, 314-322.
- 89. Tonelli, M., et al. (2016) Attraction of entomopathogenic nematodes to sugarcane root volatiles
- under herbivory by a sap-sucking insect *Chemoecology*, 59-66.
- 625 90. Rasmann, S. et al. (2011) Direct and indirect root defences of milkweed (Asclepias syriaca):
- 626 trophic cascades, trade- offs and novel methods for studying subterranean herbivory. J. Ecol. 99,
- 627 16-25.

- 91. Rasmann, S. et al. (2005) Recruitment of entomopathogenic nematodes by insect-damaged
- 629 maize roots. *Nature* 434, 732.
- 92. Aratchige, N. et al. (2004) Below-ground plant parts emit herbivore-induced volatiles: olfactory
- responses of a predatory mite to tulip bulbs infested by rust mites. Exp. Appl. Acarol. 33, 21-30.
- 93. Erb, M. et al. (2013) The role of plant primary and secondary metabolites in root-herbivore
- behaviour, nutrition and physiology. In Advances in insect physiology, pp. 53-95, Elsevier.
- 94. Hu, L. *et al.* (2018) Plant iron acquisition strategy exploited by an insect herbivore. *Science* 361, 694-697.
- 95. Penuelas, J. and Sardans, J. (2009) Ecological metabolomics. Chem. Ecol. 25, 305-309.
- 96. Heald, C. et al. (2008) Predicted change in global secondary organic aerosol concentrations in
- response to future climate, emissions, and land use change. J. Geophys. Res-Atmos. 113, D05211.
- 639 97. Wezel, A.P.v. and Opperhuizen, A. (1995) Narcosis due to environmental pollutants in aquatic
- organisms: residue-based toxicity, mechanisms, and membrane burdens. *Crit. Rev. Toxicol.* 25, 255-641 279.
- 98. Glasius, M. and Goldstein, A.H. (2016) Recent discoveries and future challenges in atmospheric
- organic chemistry. Environ. Sci. Technol. 50, 2754–2764.
- 99. Barreira, L.M.F. et al. (2015) Field measurements of biogenic volatile organic compounds in the
- atmosphere by dynamic solid-phase microextraction and portable gas chromatography-mass
- 646 spectrometry. *Atmos. Environ.* 115, 214-222.
- 647 100. Bäck, J. et al. (2010) Variable emissions of microbial volatile organic compounds (MVOCs)
- from root-associated fungi isolated from Scots pine. Atmos. Environ. 44, 3651-3659.
- 649 101. de Rijke, E. et al. (2006) Analytical separation and detection methods for flavonoids. J.
- 650 *Chromatogr. A* 1112, 31-63.
- 651 102. Weston, L.A. and Mathesius, U. (2013) Flavonoids: their structure, biosynthesis and role in the
- 652 rhizosphere, including allelopathy. J. Chem. Ecol. 39, 283-297.
- 653 103. Heath, M.C. (2000) Hypersensitive response-related death. Plant Mol. Biol. 44, 321-334.
- 654 104. Cesco, S. et al. (2010) Release of plant-borne flavonoids into the rhizosphere and their role in
- 655 plant nutrition. Plant Soil 329, 1-25.
- 656 105. Carlsen, S.C. et al. (2012) Fate in soil of flavonoids released from white clover (Trifolium
- 657 repens L.). Appl. Environ. Soil Sci. 2012, 1-10.
- 658 106. Halkier, B.A. and Gershenzon, J. (2006) Biology and biochemistry of glucosinolates. Annu.
- 659 Rev. Plant Biol. 57, 303-333.
- 660 107. Aniszewski, T. (2015) Alkaloids: chemistry, biology, ecology, and applications, Elsevier.
- 108. Klein-Junior, L.C. et al. (2016) Enlarging the bottleneck in the analysis of alkaloids: A review
- on sample preparation in herbal matrices. TrAC, Trends Anal. Chem. 80, 66-82.
- 663 109. Christensen, P. et al. (2018) Chemical fingerprinting of mobile volatile organic compounds in
- soil by dynamic headspace Thermal Desorption Gas Chromatography Mass Spectrometry.
- 665 *LCGC Europe* 31, 190-201.
- 666 110. Spietelun, A. et al. (2010) Current trends in solid-phase microextraction (SPME) fibre
- 667 coatings. Chem. Soc. Rev. 39, 4524-4537.
- 668 111. Ferreira, A.M.C. et al. (2011) Stir bar sorptive extraction of parabens, triclosan and methyl
- 669 triclosan from soil, sediment and sludge with in situ derivatization and determination by gas
- 670 chromatography—mass spectrometry. J. Chromatogr. A 1218, 3837-3844.
- 671 112. EPA, Closed system purge-and-trap extraction for volatile organics in soil and waste samples.,
- 672 US Environmental Protection Agency, 1996, pp. 1-24.

673

674

- **BOX 1.** Sampling of PSCs for identification and quantification.
- The heterogeneous chemical and physical nature of soil creates spatial differences that complicate
- 678 representative sampling. In addition, PSCs possess a high diversity of functional groups, volatilities,
- and polarities. Their belowground presence can thus be in air, in pore water, or adsorbed to solid
- 680 matter. Representative sampling may therefore require a combination of sampling techniques.
- Sampling techniques for PSCs in soil include extraction (solvent or solid phase), head-space
- analysis, and purge and trap techniques [7].

- Solvent extraction relies on the principle of "like dissolves like". Polar compounds are thus
- extracted in polar solvents (e.g. water, methanol) and nonpolar compounds in nonpolar solvents
- 685 (e.g. n-hexane, dichloromethane). Several hundreds of grams of soil, however, may be needed for
- the representative sampling of PSCs with low concentrations in soil [109]. Dilution beyond
- detection limits may occur, so subsequent sample concentration by solid-phase extraction (SPE) or
- evaporation (with the risk of loss of volatiles) may be necessary. Emission of PSCs directly from roots
- have been investigated using primarily solid-phase micro extraction (SPME) or adsorption tubes (with e.g.
- Tenax) coupled with GC-MS analysis in a number of studies as reviewed in [7].
- 691 Solid-phase micro extraction (SPME), head-space analysis, and purge and trap methods avoid the
- 692 dilution effect of solvents but are more selective in the compounds sampled. SPME relies on the
- 693 equilibrium between a fibre coated with a sorbing material and most often the gas phase (potentially
- also water or solid phase). This technique allows in-situ sampling for evaluating small-scale spatial
- differences. The sorbing material determines the selectivity of the compounds sampled [110],
- 696 whereas soil type, water content, and pH influence the equilibrium. Stir-bar sorptive extraction
- 697 (SBSE), where a magnetic rod is coated with sorbing material, may provide up to 1000-fold higher
- sensitivity but requires thermal desorption for subsequent instrument introduction. SBSE is useful
- 699 for extraction from water, as demonstrated for many organic pollutants [111]. Identifying the mode
- of transport of PSCs (i.e. in gas or water phase) could be accomplished by sampling the soil gas
- 701 phase using SPME and subsequently washing the soil and extracting using SBSE.
- 702 Collecting only head spaces using an air-tight syringe is a potential way to sample only volatile
- PSCs (i.e. VOCs). The concentration of VOCs from soil, however, may be <10 μg m⁻³, which is
- below the detection limit for head-space analysis. This limitation may be overcome with purge and
- trap methods recommended for analysis when concentrations are <200 µg kg⁻¹ [112], which have
- recently been applied for sampling isoprene, monoterpenes, and sesquiterpenes in forest soil [19].

708 **Allelochemical** is a compound produced by an organism that can have a detrimental physiological 709 effect on individuals of another species when released into the environment. 710 711 **Allelopathic effect** is the interference effect of chemical compounds released by a plant on the 712 growth and performance of other plants. An allelopathic effect usually implies a direct negative (harmful) effect, but some definitions also include positive effects. 713 714 Narcotic effect refers to the unspecific effects of lipophilic organic chemicals on membrane fluidity. These chemicals tend to partition into the lipid compartments of the organism, causing 715 716 (nonpolar) narcosis or baseline toxicity. 717 Phytotoxic refers to chemical compounds toxic to plants, including any adverse effect on plants such as delayed seed germination, inhibition of growth, death, or destruction of plant tissue. 718 719 Plant secondary compounds (PSCs) are organic compounds typically formed from primary metabolites in specific pathways. Some PSCs are involved in plant's primary metabolic processes, 720 but the large majority of these are known to serve as infochemicals mediating interactions and are 721 important for plants to survive in the environment, e.g. by their repellent, attractant, or toxic effects 722 on other organisms. 723 724 **Primary metabolites** are compounds from metabolic pathways present in all plants and are essential for growth, development, and reproduction. These metabolites comprise amino acids, 725 nucleic acids, peptides, various carbohydrates, and lipids. 726 727 Rhizosphere is the microecological zone of soil in direct proximity to plant roots and includes the 728 organisms in that soil that are directly affected by root exudates. 729 730 731 732 733

707

Glossary box:

Legend to Figure 1. Direct and indirect effects of climatic conditions (red box) and aboveground herbivore pressure (green box) on plant secondary compound production (yellow boxes) and input into the soil, and its belowground interactions (brown box). In addition, climate affects how PSCs are transported and degrade in soil. Dashed arrows and boxes indicate processes and effects that are less well understood and are elucidated in the main text.

Table 1. Overview of the important classes of PSCs involved in mediating belowground interactions, and best practices for sampling and analysing them. The chemical properties of PSCs affect their position and transport in the soil matrix, with polar PSCs dissolved in water, volatile PSCs diffusing through air spaces in the soil, and less polar non-volatile PSCs adsorbed to the organic fraction of soil.

Compound	Example of	Polarity	Water	Volatility	Ecological interactions/	The analytical instrument for detection
class	structure		solubility		impacts and effects on	in soil (see also Box 1)
					associated soil organisms	
Terpenes		Low	Low	High	Defensive and signalling	Monoterpenes and highly volatile sesquiterpenes: S
					compounds, many with a	
Synthesized					characteristic odour (e.g. α-	with PDMS adsorbent, purge-and-trap with Tenax
from isoprene	H ₂ C				pinene from pine, limonene	adsorbent and subsequent GC-MS [19], [99],
units (C_5H_8) .	H₃C				from citrus). Antioxidant,	ausorbent and subsequent GC-Wis [17], [37],
Includes	CH₃				antimicrobial, and	On-line measurement with Proton-transfer-reaction
monoterpenes					phytotoxic properties [3].	MS [100].
$(C_{10}H_{16})$ and					Toxic to invertebrates due	
sesquiterpenes					to baseline toxicity	Nonpolar sesquiterpenes likely adsorbed to soil
$(C_{15}H_{24}).$	α -Pinene				("narcotic" effect) via	
					interaction with cellular	particles require extraction with organic solvents as
					membranes [97]. Involved	evaporation followed by GC-MS [41].
					in formation of aerosol	evaporation followed by Ge 1915 [41].
					particles, thus affecting	
					climate [98].	
Flavonoids		Low to	Low to	Low	Both defensive and	Most likely present belowground in water and ac
		intermediate	intermediate		signalling; promote	Extraction using polar solvents (water, methanol)
General					symbiosis (used by legumes	metals are formed [104].
structural C-15			Presence of		to attract rhizobia), various	
skeleton	ö		sugar		antimicrobial activities [101,	Extracts can be analysed with reversed-phase LC-
consisting of			moieties		102]. Can interact with Ca ²⁺	
two phenyl rings			increases			

and a	Basic flavonoid		solubility in		channels in cell membranes	Aglycones may be sampled using SPE [101] and a
heterocyclic ring	structure		water		causing cytotoxic influx of	COMO 6 1 1 2 2 2 C
(C6-C3-C6).					Ca ²⁺ [103].	to GC-MS after derivatization. Greater sensitivity r
						obtained with SBSE and subsequent GC-MS.
Glucosinolates	O	High	High	Low	Due to polarity and water	Mostly present in the water phase due to high po
					solubility, are readily taken	Extraction using polar solvents (water, methanol)
Derived from	но "ОН				up by soil organisms, where	with LC-MS [17] or GC-MS after derivatizat
glucose and an	ŌН				they can be converted to	concentration in the sample.
amino acid.	Sinigrin				highly reactive compounds	
Most are found	Singini				such as isothiocyanate and	
in the order					thiocyanates. Toxicity	
Brassicales (30).					pathways include inhibition	
					of cytochrome P ₄₅₀	
					detoxification complex,	
					induction of programmed	
					cell death (apoptosis), and	
					genotoxic effects [106].	
Alkaloids		Low to	Low to	Low	Mostly defensive	Low concentration in soil, so clean-up and pr
		intermediate	intermediate		compounds. Phytotoxic,	necessary. Polarity can be adjusted with pH changes
Highly diverse,	\sim H \sim				toxic (e.g. strychnine), and	with liquid-liquid extraction, SPE, or SPME [108]
usually with a	() N				narcotic effects (e.g.	both LC-MS and GC-MS.
basic nitrogen	N				morphine, codeine). Can	
atom either in	Nicotine				interfere with gene	

cyclic rings or		transcription and protein	
open chains.		synthesis [107].	

Outstanding Questions

To what extent does the plant secondary compound (PSC) plume originally released into a soil compartment become chemically degraded during diffusion and mixed with secondarily formed and other ambient PSCs? Does this dilution and mixing with other PSCs impair the signal quality for the receiving soil organisms or have they adapted to and even benefited from signal noise, e.g. to improve their foraging efficiency?

How will climate change affect PSC production in both above- and belowground organs and how will these effects in turn affect the belowground interactions of associated communities of microbes, plants, and soil invertebrates?

What is the relative importance of specific PSCs for both plant-soil invertebrate and plant-plant interactions and can we understand their prevalence from knowledge on their mode of action? If the effects of PSC on associated species is very compounds specific, quantitative and qualitative changes in the input of PSC to the soil can have cascading effects on richness and composition of belowground soil communities that may ultimately alter soil functioning.

What is the importance of a common co-evolutionary history between PSCs in soil and the organisms exposed to them for shaping the outcome of their interactions?

How important are belowground interactions in determining the composition and concentration of PSCs released by the producer? In particular, what is the impact of selection (if any) that belowground interactions can exert on PSC composition and production in leaves relative to aboveground interactions and climate?

HIGHLIGHTS:

- Plants produce a high diversity of secondary compounds that are released to their surroundings.
- The role of plant secondary compounds (PSCs) on aboveground interactions has been well studied, but less is known about the role of PSCs in mediating and determining the outcome of belowground interactions especially interactions beyond the rhizosphere.
- PSCs enter the soil directly by excretion from belowground plant organs or indirectly by washing from the canopy and leaching from litter.
- Leaching of PSCs creates a heterogeneous soil chemical environment and mediate plant-plant interactions, plant-microbiome associations, and plant-soil invertebrate relationships. This affects and shape the diffuse community of species living in soil also further away from the rhizosphere.
- Climate change can substantially alter the quantitative and qualitative production of PSCs released to the soil, which may change belowground species interactions and affect species composition, richness, and ultimately soil functioning.

